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# Variation Among Species and Populations, and Carry-Over Effects of Winter Exposure on Mercury Accumulation in Small Petrels

Petra Quillfeldt<sup>1\*</sup>, Yves Cherel<sup>2</sup>, Joan Navarro<sup>3</sup>, Richard A. Phillips<sup>4</sup>, Juan F. Masello<sup>1</sup>, Cristián G. Suazo<sup>1</sup>, Karine Delord<sup>2</sup> and Paco Bustamante<sup>5,6</sup>

<sup>1</sup> Department of Animal Ecology and Systematics, Justus Liebig University Giessen, Giessen, Germany, <sup>2</sup> Centre d'Etudes Biologiques de Chizé, UMR 7372 CNRS - La Rochelle Université, Villiers-en-Bois, France, <sup>3</sup> Institut de Ciències del Mar, Consejo Superior de Investigaciones Científicas, Barcelona, Spain, <sup>4</sup> British Antarctic Survey, Natural Environment Research Council, Cambridge, United Kingdom, <sup>5</sup> Littoral Environnement et Sociétés, UMR 7266 CNRS - La Rochelle Université, La Rochelle, France, <sup>6</sup> Institut Universitaire de France, Paris, France

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### \*Correspondence:

Petra Quillfeldt  
petra.quillfeldt@bio.uni-giessen.de

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Even in areas as remote as the Southern Ocean, marine organisms are exposed to contaminants that arrive through long-range atmospheric transport, such as mercury (Hg), a highly toxic metal. In previous studies in the Southern Ocean, inter-specific differences in Hg contamination in seabirds was generally related to their distribution and trophic position. However, the Blue Petrel (*Halobaena caerulea*) was a notable exception among small seabirds, with higher Hg levels than expected. In this study, we compared the Hg contamination of Blue Petrels and Thin-billed Prions (*Pachyptila belcheri*), which both spend the non-breeding season in polar waters, with that of Antarctic Prions (*Pachyptila desolata*), which spend the winter in subtropical waters. We collected body feathers and blood samples, representing exposure during different time-frames. Hg concentrations in feathers, which reflect contamination throughout the annual cycle, correlated with  $\delta^{13}\text{C}$  values, and varied with ocean basin and species. Blue Petrels from breeding colonies in the southeast Pacific Ocean had much higher feather Hg concentrations than expected after accounting for latitude and their low trophic positions. Both Hg concentrations and  $\delta^{15}\text{N}$  in blood samples of Blue Petrels were much lower at the end than the start of the breeding period, indicating a marked decline in Hg contamination and trophic positions, and the carry-over of Hg burdens between the wintering and breeding periods. Further parameters such as myctophids as prey and foraging in the sea-ice environment may lead to elevated Hg levels. Our study underlines that carry-over of Hg concentrations in prey consumed in winter may determine body Hg burdens well into the breeding season.

**Keywords:** distribution, mercury, petrels, stable isotopes, trophic position

## INTRODUCTION

Seabirds are often used as sentinels of marine pollution (Van den Steen et al., 2011; Becker et al., 2016; Thébaud et al., 2021). They are long-lived animals, feed at high trophic levels, and thus integrate and bioaccumulate contaminants from the food webs on which they rely (Albert et al., 2019). Often, seabirds nest in accessible breeding colonies, but roam over vast areas of ocean that

115 can thus be monitored. Our knowledge of their diets and at- 172  
116 sea distribution has greatly increased in the last years with the 173  
117 advances in biologging methods that are now suitable for the 174  
118 smallest seabird species (Quillfeldt et al., 2015), trophic tracers 175  
119 such as compound-specific stable isotope analyses (Lorrain et al., 176  
120 2009; Quillfeldt and Masello, 2020), and metabarcoding from 177  
121 faecal samples (Kleinschmidt et al., 2019). 178

122 Among the contaminants that increase in the marine 179  
123 environment due to human activities, mercury (Hg) is a highly 180  
124 toxic non-essential metal that has deleterious effects on the 181  
125 behaviour, neurology, endocrinology and development of wildlife 182  
126 (Scheuhammer et al., 2007; Tan et al., 2009). Released from both 183  
127 natural and anthropogenic sources, Hg reaches remote polar 184  
128 and sub-polar regions through long-range atmospheric transport 185  
129 (Fitzgerald et al., 1998). In seabirds, Hg is incorporated from the 186  
130 food and accumulates in soft tissues such as liver and muscle 187  
131 (Bearhop et al., 2000a; Carravieri et al., 2014a). Birds can excrete 188  
132 up to 90% of the Hg accumulated since the previous moult 189  
133 in the new growing feathers and thus, feathers – which can 190  
134 be sampled non-destructively – are an archive of year-round 191  
135 Hg contamination (Thompson et al., 1998; Albert et al., 2019). 192  
136 Birds may also show a substantial carry-over of Hg among 193  
137 seasons, and slow changes in Hg over time. For example, Double- 194  
138 Crested Cormorants (*Phalacrocorax auritus*) and Caspian Terns 195  
139 (*Hydroprogne caspia*) with high Hg exposure in winter still had 196  
140 elevated blood Hg values in summer (Lavoie et al., 2014). 197

141 Among seabirds, species with high trophic position in 198  
142 marine food webs have elevated Hg concentrations due to 199  
143 the biomagnification of methylmercury (MeHg), the most 200  
144 bioavailable form of Hg in marine ecosystems (Seco et al., 2021). 201  
145 This pattern has been shown in the seabird community of the 202  
146 subantarctic Kerguelen Islands (Blévin et al., 2013; Carravieri 203  
147 et al., 2014a). In particular, species feeding in colder waters to 204  
148 the south had lower Hg concentrations than species feeding 205  
149 in northern, warmer waters. At the scale of the Southern 206  
150 Hemisphere, such a pattern (higher Hg concentrations in 207  
151 birds feeding in subtropical and subantarctic waters) has been 208  
152 confirmed for diverse species, including penguins, skuas, and 209  
153 albatrosses (Carravieri et al., 2014b, 2016, 2017, 2020; Cherel 210  
154 et al., 2018). However, the Blue Petrel (*Halobaena caerulea*) 211  
155 seems to be a marked exception to this general pattern, as Hg 212  
156 concentrations in tissues is one order of magnitude higher than 213  
157 in other species of small petrels (Bocher et al., 2003). 214

158 The Blue Petrel is a similar size (~200 g) to Prions, *Pachyptila* 215  
159 spp. The largest breeding populations are at Diego Ramírez 216  
160 Islands, Chile in the southeast Pacific Ocean (>2 million 217  
161 individuals or ~1.35 million pairs; Schlatter and Riveros, 1997; 218  
162 Lawton et al., 2006), Kerguelen Islands in the southern Indian 219  
163 Ocean (100,000–200,000 pairs; Weimerskirch et al., 1989) and 220  
164 Marion Island in the Indian Ocean (110,000–180,000 pairs; Dilley 221  
165 et al., 2017). Muscle tissue sampled from Blue Petrels breeding 222  
166 at Kerguelen Islands contains far higher Hg concentrations 223  
167 than expected, given the relatively low Hg levels in epipelagic 224  
168 fish and crustaceans in the same region (Bocher et al., 2003). 225  
169 Proposed explanations include the relative longevity of Blue 226  
170 Petrels (up to 20 years) and thus, Hg bioaccumulation over 227  
171 the long-term, and from their consumption of mesopelagic fish 228

(Cherel et al., 2002b), which contain high Hg concentrations 172  
(Bustamante et al., 2003; Cipro et al., 2018; Seco et al., 2020). 173  
Blue Petrels at Marion Island in the southern Indian Ocean 174  
showed the highest feather Hg concentrations reported for 175  
the species so far (Becker et al., 2016). At South Georgia, 176  
studies reported either relatively high Hg concentrations in 177  
feathers of Blue Petrels (Becker et al., 2002) or Hg levels 178  
in a similar range to Antarctic prions and diving petrels 179  
(Anderson et al., 2009). 180

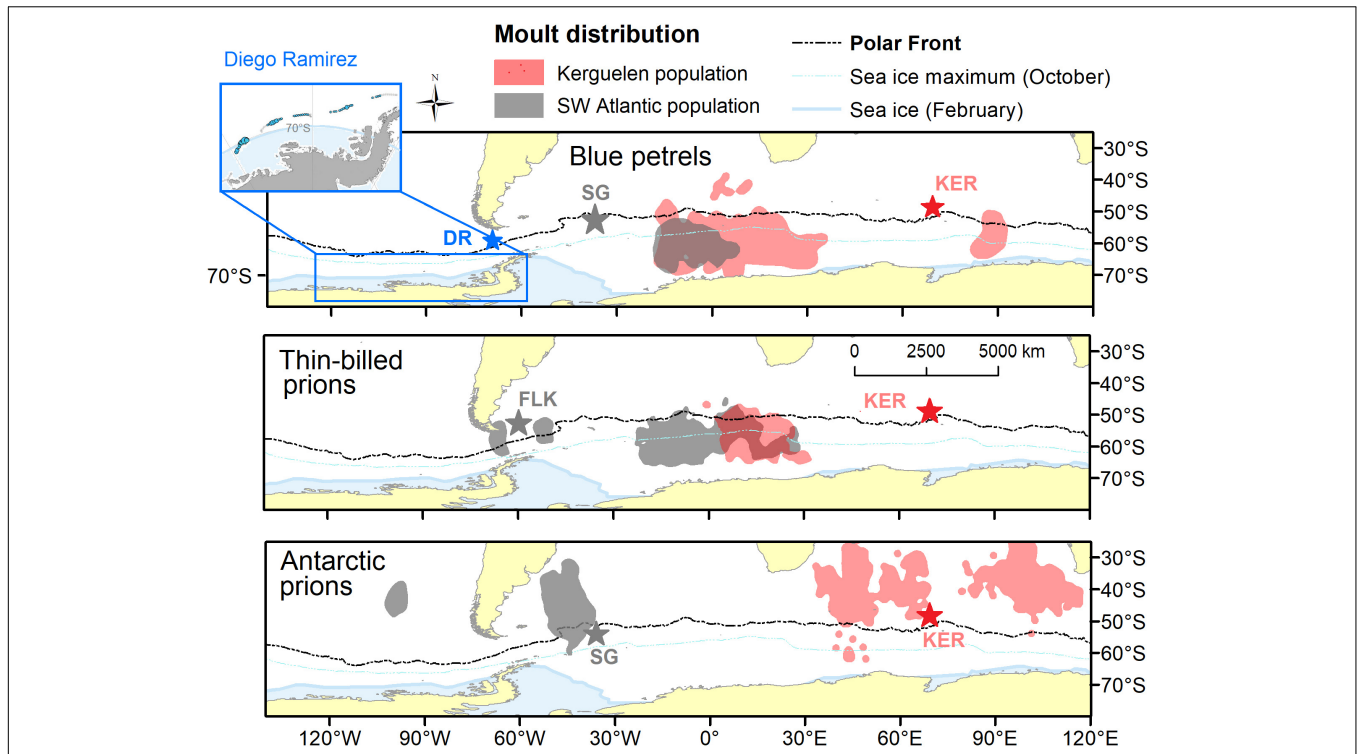
181 Although Hg in Southern Ocean seabirds has received 182  
considerable attention (Anderson et al., 2009; Becker et al., 183  
2016; Blevin et al., 2017; Carravieri et al., 2020), the influence 184  
of sea ice on Hg dynamics has not yet been explored. Recent 185  
studies identified bacteria of the genus *Nitrospina* as a potential 186  
Hg methylator within sea ice and brine, and proposed that 187  
Antarctic waters associated with sea ice can harbour a microbial 188  
source of MeHg in the Southern Ocean (Gionfriddo et al., 189  
2016). Thus, total Hg (i.e., inorganic Hg and MeHg) and 190  
methylated Hg (MeHg) concentrations are elevated in these 191  
zones, related to high atmospheric Hg deposition and subsequent 192  
*in situ* methylation (Gionfriddo et al., 2016). A study of the 193  
Hg species distribution suggested that the Southern Ocean 194  
Hg cycle is characterized by a net atmospheric Hg deposition 195  
on surface waters near the ice edge, and Hg enrichment in 196  
brine during sea-ice formation (Cossa et al., 2011). Studies 197  
in coastal Antarctica have shown greatly enhanced total Hg 198  
concentrations in surface snow at the sea-ice edge adjacent to the 199  
freezing ocean surface (McMurdo/Ross Sea region: Brooks et al., 200  
2008; Casey station/East Antarctic: Cossa et al., 2011). The Hg 201  
concentrations found in fast ice near Casey station were three 202  
orders of magnitude above the concentrations in surface water 203  
in the Southern Ocean (Cossa et al., 2011). A seasonal study of 204  
elemental and total Hg concentrations in the Antarctic sea-ice 205  
environment (Nerentorp Mastromonaco et al., 2016) found that 206  
the concentration of total Hg in sea ice halved from winter to 207  
spring (average 9.7 ng/l to 4.7 ng/l). A recent analysis has related 208  
high winter Hg concentrations to the frequency of katabatic 209  
winds, bringing Hg from the Antarctic ice sheet to coastal waters 210  
(Yu et al., 2021).

211 In the present study, we compared Hg concentrations in 212  
blood and feathers of Blue Petrels, Antarctic Prions (*Pachyptila* 213  
*desolata*), and Thin-billed Prions (*P. belcheri*), each at their largest 214  
colonies in widely separated oceans. Of the three species, Blue 215  
Petrels spend the non-breeding season at the most southerly 216  
latitudes (Quillfeldt et al., 2013, 2015; Navarro et al., 2015), and 217  
have disproportionately high Hg values (Bocher et al., 2003). We 218  
therefore used tracking data to examine if exposure to sea ice may 219  
play a part in explaining variability in Hg concentrations. We 220  
used stable isotope analyses to determine trophic positions and 221  
distributions (water mass) used by each species. In the Southern 222  
Ocean,  $\delta^{13}\text{C}$  values in seabird tissues correspond to the location 223  
of their foraging habitats (Phillips et al., 2009; Jaeger et al., 2010; 224  
Quillfeldt et al., 2010b) and  $\delta^{15}\text{N}$  values increase with trophic 225  
position (Cherel et al., 2010). As novel questions, we aimed to 226  
test (1) if foraging close to sea-ice-covered polar waters results 227  
in higher exposure to Hg, and (2) if there is carry-over of Hg 228



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**FIGURE 1 |** Distribution of Blue Petrels, Thin-billed Prions, and Antarctic Prions during primary moult (i.e., the core moult area). Colony sites: Diego Ramirez (DR), South Georgia (SG), Falkland Islands (Malvinas) (FLK), and Kerguelen (KER). Blue Petrels moulting in February between 71°S, 119°W and 67°S, 78°W are most likely birds from the large Diego Ramirez colony (Ryan et al., 2020). Moult takes place around the time of the minimum sea-ice extent (February–April) in Blue Petrels and Thin-billed Prions, and during August–October in Antarctic Prions.

## MATERIALS AND METHODS

### Study Species

Blue Petrels, Thin-billed Prions, and Antarctic Prions have wide distributions in the Southern Ocean. We sampled breeding populations in the south-west Atlantic Ocean (Falkland Islands for Thin-billed Prions; South Georgia for Blue Petrels and Antarctic Prions) and in the Indian Ocean (Kerguelen Islands, all three species) (Figure 1). In addition, a population of Blue Petrels was sampled on Diego Ramirez Islands, Chile, southeast Pacific Ocean. In total, we sampled seven populations (Figure 1). Thin-billed Prions breed mainly on the Falkland and Kerguelen Islands. New Island, in the Falkland Islands, is the most important known breeding site for Thin-billed Prions with an estimated two million breeding pairs. South Georgia and Kerguelen are the most important breeding sites (with populations > 1 million) of Antarctic Prions.

These three petrel species migrate away from their breeding grounds during the non-breeding season, where they segregate latitudinally (Navarro et al., 2015; Quillfeldt et al., 2015). Antarctic Prions migrate to subtropical waters, and Thin-billed Prions and Blue Petrels moult in polar waters (Quillfeldt et al., 2013, 2015; Navarro et al., 2015). The species also show breeding allochrony, with Blue Petrels arriving at colonies in September, Thin-billed Prions in October and Antarctic Prions in November

to early December (Quillfeldt et al., 2020). After several days of pair formation, the birds leave on a pre-laying exodus, and return ready for egg-laying and incubation, with the mean start of the first trip by the female in incubation at Kerguelen of 28 October (Blue Petrel), 19 November (Thin-billed Prion) and 26 December (Antarctic Prion) (Quillfeldt et al., 2020).

Differences in habitat use in the breeding season are less pronounced than in winter, and diets largely overlap. The three species are zooplanktivorous, with a preference for crustaceans (Prince, 1980; Cherel et al., 2002a,b; Quillfeldt et al., 2010a), and forage on the surface or up to depths of 5–7 m (Chastel and Bried, 1996; Cherel et al., 2002a; Navarro et al., 2013).

### Study Sites and Seasons

Adult Blue Petrels and the two species of Prions were trapped either at the burrow or by mist net. Fieldwork at Kerguelen was carried out in colonies of Thin-billed Prions and Blue Petrels at Île Mayès (49°28'S, 69°57'E) during incubation, late chick-rearing or post-moult periods (when Blue Petrels return to clean out their burrows) of five breeding seasons (Tables 1, 2). Sampling in 2010/11 was carried out as part of the POLARTOP project (Carravieri et al., 2014a,b) and in 2011/12, blood and feather samples were collected during the deployment and retrieval of geolocator-immersion loggers (Quillfeldt et al., 2015). Antarctic Prions were sampled at Île Verte (49°30'S, 70°02'E; n = 10) in 2011/12. Mist netting of Blue Petrels was carried

**TABLE 1** | Summary of stable isotope and mercury data of Blue Petrels (mean ± standard deviation).

	<b>POLARTOP Kerguelen 2010/11</b>	<b>GLS deployments Kerguelen 2011/12</b>	<b>GLS recoveries Kerguelen 2012/13</b>	<b>Kerguelen 2018/19</b>	<b>Diego Ramirez 2010/11</b>	<b>South Georgia 2010/11</b>	<b>GLS recoveries South Georgia 2011/12</b>
<b>Body feathers</b>							
N	10	Not sampled	17	20	30 (16 for Hg)	20	8
δ <sup>13</sup> C (‰)	-24.4 ± 0.7		-24.9 ± 0.5	-25.7 ± 1.1	-23.6 ± 1.3	-25.0 ± 1.3	-24.8 ± 0.8
δ <sup>15</sup> N (‰)	9.0 ± 0.4		8.6 ± 0.5	8.3 ± 0.5	10.3 ± 0.9	8.8 ± 0.9	9.1 ± 0.7
TP <sub>CSIA</sub>				3.21 ± 0.04	3.79 ± 0.11		
TP <sub>LM</sub>				3.27 ± 0.05	3.43 ± 0.07		
Hg (μg/g dw)	1.44 ± 0.42		2.09 ± 1.65	1.68 ± 0.96	4.42 ± 2.72	1.69 ± 1.51	1.09 ± 0.72
<b>Blood (early breeding season)</b>							
N (sample time)	10 (September)	Not sampled	17 (November)	20 (November)	Not sampled	16 (20 November–4 December)	Not sampled
δ <sup>13</sup> C (‰)	-22.4 ± 1.2		-24.0 ± 0.9	-24.1 ± 0.9		-23.4 ± 0.6	
δ <sup>15</sup> N (‰)	10.3 ± 0.8		9.3 ± 0.5	9.2 ± 0.6		9.7 ± 0.4	
TP <sub>CSIA</sub>	-		-	3.55 ± 0.24		-	
Hg (μg/g dw)	6.00 ± 2.78		4.58 ± 1.83	4.01 ± 1.63		2.76 ± 1.81	
<b>Blood (late breeding season)</b>							
N	11 (February)	20 (29 December–6 January)	Not sampled	20 (April)	24 (6 December–26 January)	Not sampled	Not sampled
δ <sup>13</sup> C (‰)	-24.3 ± 0.5	-23.9 ± 1.1		-27.0 ± 0.3	-24.6 ± 0.2		
δ <sup>15</sup> N (‰)	8.0 ± 0.3	9.1 ± 0.3		7.9 ± 0.4	8.8 ± 0.5		
TP <sub>CSIA</sub>	-	-		-	3.43 ± 0.06		
Hg (μg/g dw)	2.06 ± 0.74	2.43 ± 1.05		0.49 ± 0.15	2.92 ± 0.74		

Early breeding season: arrival (September) to incubation (November), late breeding season: chick-feeding (December–February) to post-moult return (April).

**TABLE 2** | Summary of stable isotope and mercury data of Thin-billed Prions (mean ± standard deviation).

	<b>New Island Falkland/Malvinas 2006/07</b>	<b>Falkland/Malvinas 2017/18</b>	<b>POLARTOP Kerguelen 2010/11</b>	<b>GLS recoveries Kerguelen 2012/13</b>	<b>Kerguelen 2018/19</b>
<b>Feathers (moult)</b>					
N	20	20	12	23	14
δ <sup>13</sup> C (‰)	-22.1 ± 2.8	-21.6 ± 1.8	-24.0 ± 1.0	-23.5 ± 1.0	-25.3 ± 0.9
δ <sup>15</sup> N (‰)	10.5 ± 3.4	10.7 ± 1.94	9.1 ± 0.3	8.7 ± 0.3	8.2 ± 0.4
TP <sub>CSIA</sub>	3.53 ± 0.06	3.39 ± 0.10			3.34 ± 0.07
TP <sub>LM</sub>	3.51 ± 0.27	3.50 ± 0.14			3.27 ± 0.03
Hg (μg/g dw)	0.76 ± 0.61	1.13 ± 0.74	0.90 ± 0.29	1.62 ± 0.67	1.04 ± 0.52
<b>Blood (early breeding season)</b>					
N (month)	12	20	10 (October)	23 (26 November–3 December 2012)	14 (November)
δ <sup>13</sup> C (‰)	-18.8 ± 0.8	-19.8 ± 0.5	-23.4 ± 1.5	-23.3 ± 1.2	-23.8 ± 0.5
δ <sup>15</sup> N (‰)	12.4 ± 1.2	11.2 ± 1.1	9.3 ± 0.6	8.9 ± 0.3	8.2 ± 0.3
TP <sub>CSIA</sub>		3.60 ± 0.07			3.56 ± 0.08
Hg (μg/g dw)	0.80 ± 0.25	0.99 ± 0.25	1.46 ± 0.39	1.29 ± 0.39	1.31 ± 0.31
<b>Blood (late breeding season)</b>					
N	6	20	12 (February)	Not sampled	3 (April)
δ <sup>13</sup> C (‰)	-19.5 ± 1.9	-17.9 ± 1.1	-24.0 ± 0.6		-25.1 ± 0.2
δ <sup>15</sup> N (‰)	12.1 ± 1.3	11.9 ± 0.9	8.0 ± 0.2		7.5 ± 0.2
TP <sub>CSIA</sub>		3.47 ± 0.05			
Hg (μg/g dw)	0.61 ± 0.24	0.63 ± 0.15	0.73 ± 0.20		0.72 ± 0.12

Early breeding season: arrival (October) to incubation (December), late breeding season: chick-feeding (January–April).

457 out at Isla Gonzalo, Diego Ramírez Islands (56°29'S, 68°44'W)  
 458 in December 2010 to January 2011. Thin-billed Prions were  
 459 sampled at New Island, Falkland/Malvinas Islands (51°43'S,  
 460 61°18'W) in 2006/07 and 2017/18. Blue Petrels and Antarctic  
 461 Prions were sampled at Bird Island, South Georgia (54°00'S,  
 462 38°03'W) in burrows during the austral summer 2010/11, when  
 463 the incubation period overlaps between the two species, and  
 464 feathers were also collected from Blue Petrels when geolocators  
 465 were retrieved in austral summer 2011/12.

466 **Sample Collection**

467 We sampled two different tissue types, body feathers and blood.  
 468 Body feathers, moulted annually, represent Hg accumulated over  
 469 the annual cycle (Albert et al., 2019). To assess seasonal changes  
 470 in Hg exposure, we sampled blood at different stages in the  
 471 breeding season as blood reflects the contamination for the 1–2  
 472 previous months (half-life of 30 days in Great Skuas *Stercorarius*  
 473 *skua*: Bearhop et al., 2000; 40–65 days in Cory's shearwaters  
 474 *Calonectris borealis*: Monteiro and Furness, 2001). For sample  
 475 times and sizes see **Tables 1–3**.

476 Feather samples (body feathers) were stored in individual  
 477 Ziploc bags. Antarctic Prions moult their primaries towards  
 478 the end of the non-breeding season, and Blue Petrels and  
 479 Thin-billed Prions directly after the breeding season (Cherel  
 480 et al., 2016). Less is known about body feather moult,  
 481 but this is thought to occur over a longer period. Blue  
 482 Petrels collected in January (i.e., likely non-breeders or failed  
 483 breeders) had extensive body moult coinciding with primary  
 484 and secondary feather moult (Bierman and Voous, 1950), but  
 485 very few Blue Petrels moult body feathers in winter (Brown  
 486 et al., 1986). Blue Petrels return to the colony after their  
 487  
 488  
 489

514 moult, mostly in May (Brooke et al., 2004; own observations  
 515 from tracking data).

516 Feathers were cleaned in a chloroform:methanol solution (2:1,  
 517 v/v) in an ultrasonic bath and rinsed two times in methanol. After  
 518 48 h drying at 45°C in an oven, they were cut into tiny fragments  
 519 with stainless steel scissors. Blood (0.2–0.4 ml) was sampled  
 520 by puncture of the wing vein and collected using heparinized  
 521 capillaries, or syringes. Blood was stored in ethanol (Diego  
 522 Ramírez, Kerguelen 2012/13), or separated by centrifugation, and  
 523 the pellet of red blood cells was frozen (Kerguelen 2010/11 and  
 524 2018/19, Falkland Islands, and South Georgia). Both whole blood  
 525 and blood cells were freeze-dried and ground to powder for Hg  
 526 and stable isotope analyses. As Hg from whole blood is mainly  
 527 found in red blood cells (>95%), it is equivalent to analyse one or  
 528 the other, when referring to dry mass.

529 The half-life of isotope turnover for avian red blood cells was  
 530 29.8 days in crows (*Corvus brachyrhynchos*) (Hobson and Clark,  
 531 1993). For this, blood samples collected from petrels therefore  
 532 likely represented the diet ingested ca. 2–4 weeks before sampling.  
 533 After return from the wintering areas, stable isotope ratios in  
 534 blood quite quickly reach values characteristic of the summer  
 535 habitat and diet (Cherel et al., 2014; Lavoie et al., 2014). In  
 536 contrast, there can be substantial carry-over of Hg among seasons  
 537 and slow changes in the body pool of Hg over time, especially  
 538 for individuals with high Hg exposure in winter (Lavoie et al.,  
 539 2014). This suggests a slow depuration rate and storage in internal  
 540 tissues, such that levels in the blood reflect both recent and past  
 541 exposure. Renal excretion of MeHg is low and bile excretion  
 542 is followed by intestinal reabsorption, thus retaining Hg in the  
 543 organism. Hence, Hg values in blood at a given time may be  
 544 influenced by previous exposure at distant locations.  
 545  
 546

490 **TABLE 3 |** Summary of stable isotope and mercury data of Antarctic Prions (mean ± standard deviation).

	Antarctic prion - GLS recoveries Kerguelen 2012/13	Antarctic prion South Georgia 2010/11	Antarctic prion - GLS recoveries South Georgia 2011/12
<b>Feathers (moult)</b>			
N	10	20	6
δ <sup>13</sup> C (‰)	-18.8 ± 0.9	-18.7 ± 1.1	-20.9 ± 1.0
δ <sup>15</sup> N (‰)	9.9 ± 0.8	10.5 ± 1.8	10.1 ± 1.0
TP <sub>LM</sub>			
Hg (μg/g dw)	2.39 ± 0.58	1.68 ± 0.75	1.49 ± 0.44
<b>Blood (early breeding season)</b>			
N (month)	10 (January)	15 (December–January)	Not sampled
δ <sup>13</sup> C (‰)	-23.8 ± 0.8	-21.8 ± 0.7	
δ <sup>15</sup> N (‰)	8.2 ± 0.2	8.2 ± 0.4	
TP <sub>LM</sub>			
Hg (μg/g dw)	0.71 ± 0.18	0.39 ± 0.13	
<b>Blood (late breeding season)</b>			
N	Not sampled	2 (February)	Not sampled
δ <sup>13</sup> C (‰)		-21.6 ± 1.8	
δ <sup>15</sup> N (‰)		8.9 ± 0.3	
TP <sub>LM</sub>			
Hg (μg/g dw)		0.34 ± 0.20	

513 *Early breeding season: incubation (December–January), late breeding season: chick-feeding (February).*



## Mercury Analyses

Mercury concentrations were determined on aliquots with an Advanced Mercury Analyser spectrophotometer Altec AMA-254 [aliquots: blood ~2 mg dry weight (dw), feathers ~1 mg dw] as described in Bustamante et al. (2006). AMA measures total Hg but bird blood and feathers contain virtually 100% methylmercury (Thompson and Furness, 1989; Renedo et al., 2017; Manceau et al., 2021). Measurements were repeated two to three times for each sample, until the relative standard deviation (RSD) was <10%. For each set of samples, accuracy and reproducibility of the results were tested by preparing analytical blanks and performing replicate measurements of certified reference materials (TORT-2: lobster hepatopancreas, certified concentration:  $0.27 \pm 0.06 \mu\text{g/g dw}$ ; DOLT-5: dogfish liver, certified concentration:  $0.44 \pm 0.18 \mu\text{g/g dw}$ ; National Research Council of Canada). Measured Hg concentrations for the certified reference materials were:  $0.26 \pm 0.02 \mu\text{g/g dw}$  ( $n = 18$ ) and  $0.42 \pm 0.01 \mu\text{g/g dw}$  ( $n = 15$ ) for TORT-2 and DOLT-5, respectively, corresponding to a recovery rate of  $96 \pm 2\%$  for TORT-2 and  $96 \pm 1\%$  for DOLT-5. The limit of detection (LOD) was  $0.005 \mu\text{g/g dw}$ . Hg concentrations are expressed in  $\mu\text{g/g dw}$ .

## Bulk Stable Isotope Analyses

To perform bulk stable isotope analyses, 0.2–0.4 mg of sample was weighed into tin cups.  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values were determined with a continuous-flow mass spectrometer (Thermo Scientific Delta V Advantage) coupled to an elemental analyser (Thermo Scientific Flash EA 1112). Results are expressed in parts per thousand (‰) in the usual  $\delta$  notation, relative to Vienna Pee Dee Belemnite for  $\delta^{13}\text{C}$  and atmospheric  $\text{N}_2$  for  $\delta^{15}\text{N}$ , following the formula:

$$\delta^{13}\text{C} \text{ or } \delta^{15}\text{N} = \left( \frac{R_{\text{sample}}}{R_{\text{standard}}} - 1 \right) \times 10^3$$

where  $R$  is  $^{13}\text{C}/^{12}\text{C}$  or  $^{15}\text{N}/^{14}\text{N}$ , respectively. Measurements of internal laboratory standards were conducted using acetanilide and peptone and indicated an experimental precision of  $\pm 0.15\%$  for both elements.

## Compound-Specific Isotope Analyses of Amino Acids

Compound-specific isotope analyses of amino acids (CSIA-AA) data can provide a good estimate of the trophic position of marine organisms even from temporally and spatially variable environments. CSIA-AA were performed at the UC Davis Stable Isotope facility (United States), as described previously (Quillfeldt and Masello, 2020). Trophic positions (TP) were calculated from the  $\delta^{15}\text{N}$  values of glutamic acid (Glx) and phenylalanine (Phe), using a stepwise trophic discrimination factor (multi-TDF<sub>Glx–Phe</sub>, for detailed discussion, see Quillfeldt and Masello, 2020), with the following equations:

$$\text{TP[feathers]} = 2 + \frac{\text{Glx} - \text{Phe} - 3.5\text{‰} - 3.4\text{‰}}{6.2\text{‰}}$$

$$\text{TP[blood cells]} = 2 + \frac{\text{Glx} - \text{Phe} - 4.0\text{‰} - 3.4\text{‰}}{6.2\text{‰}}$$

Due to high analytical costs, only small sample sizes were analysed with CSIA-AA. For Blue Petrels (Table 1), we analysed 10 blood samples and 10 feathers (five from Kerguelen and five from Diego Ramírez, respectively). For Thin-billed Prions (Table 2), we included 20 blood samples (5 from Kerguelen and 15 from New Island: 5 each in 2 years and 2 parts of the season), and 21 feathers (5 from Kerguelen and 16 from New Island: 5 from 2017 to 2018, and 11 from 2006 to 2007).

## Calculation of Trophic Positions

Trophic positions were calculated as described in Thébaud et al. (2021). In the Southern Hemisphere, a latitudinal enrichment in  $\delta^{15}\text{N}$  baseline values occurs from Antarctic to subtropical waters (Jaeger et al., 2010; Quillfeldt et al., 2010b). To correct for this latitudinal effect, we calculated the trophic positions of the birds by applying linear regression models to the relationship between TP<sub>CSIA</sub> and bulk stable isotope values ( $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$ ). Trophic positions calculated with linear models are referred as TP<sub>LM</sub>.

Linear regression models were used to test relationships between TP<sub>CSIA</sub> and bulk stable isotope values ( $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$ ). Models were applied separately for blood samples, both reflecting short-term food intake and with similar TDF – 4.0‰ (Quillfeldt and Masello, 2020) and 4.1‰ (Hebert et al., 2016), respectively, and feather samples (which reflect trophic ecology at the time of moult). For feathers, the linear regression model was statistically significant ( $R^2 = 0.58$ ,  $F_{28,2} = 19.1$ ,  $p < 0.001$ ), and the following equation was used to calculate trophic positions from bulk stable isotope values:

$$\text{TP}_{\text{LM}}[\text{feathers}, N = 31] \\ = 3.476 + 0.026 \times \delta^{13}\text{C} + 0.055 \times \delta^{15}\text{N}$$

However, the linear regression model was not statistically significant for blood TP<sub>CSIA</sub> values ( $R^2 = 0.05$ ,  $F_{22,2} = 0.5$ ,  $p = 0.596$ ). Thus, we did not calculate trophic positions from bulk stable isotope values for blood.

## Distribution, Moult and Sea Ice Concentrations

Moult times and distributions were determined using three steps, as described previously in Cherel et al. (2016): using the information recorded by the geolocator-immersion loggers (i) extraction of daily data on activity using the ACTAVE tool (Mattern et al., 2015), (ii) fitting a Generalized Additive Model (GAM) to the variable ‘on-water’ (i.e., the total time spent on water) separately for each individual, and (iii) calculating the dates when the fitted ‘on-water’ value exceeded 75% of the maximum (which indicates the core moult area; Cherel et al., 2016).

We defined habitat zones following Cherel et al. (2018), based on feather  $\delta^{13}\text{C}$  isoscapes (Jaeger et al., 2010), as Subtropical Zone (STZ):  $\delta^{13}\text{C} > -18.3\text{‰}$ , Subantarctic Zone (SAZ):  $\delta^{13}\text{C}$  values of  $-21.2$  to  $-18.3\text{‰}$ , and Antarctic Zone (AZ):  $\delta^{13}\text{C} < -21.2\text{‰}$ . Likewise, in blood, habitat was derived from  $\delta^{13}\text{C}$  as Subtropical Zone (STZ):  $\delta^{13}\text{C} > -20.1\text{‰}$ , Subantarctic Zone (SAZ):  $\delta^{13}\text{C}$  values of  $-22.9$  to  $-20.1\text{‰}$ , and Antarctic Zone (AZ):  $\delta^{13}\text{C} < -22.9\text{‰}$  (Jaeger et al., 2010).

685 The populations were assigned to the ocean basin where they  
 686 spend most of their annual cycle. Thus, although Blue Petrels  
 687 from Kerguelen moult in the Atlantic, and Blue Petrels from  
 688 South Georgia spend 2 months in winter in the Pacific, they were  
 689 assigned to the ocean basin of their breeding colony, i.e., Indian  
 690 and Atlantic Ocean, respectively.

691 Using geolocator data, we calculated an index of sea-ice  
 692 concentrations used by tracked birds, obtained through the  
 693 Environmental Data Automated Track Annotation System (Env-  
 694 DATA) on Movebank<sup>1</sup>. Sea-ice values (ECMWF Interim Full  
 695 Daily SFC Sea Ice Cover, scale 0–1) for each location were  
 696 summarized by individual and month. From these, we calculated  
 697 the maximum value and mean annual sea-ice concentration. The  
 698 maximum values were reached in the weeks before the breeding  
 699 season, and we tested for a relationship with Hg values in blood  
 700 collected in the early breeding season. An exception was the  
 701 Thin-billed Prions from New Island, where the sea-ice maximum  
 702 was reached earlier in the winter; however, this population was  
 703 excluded from analyses as Hg was not measured in feathers and  
 704 blood of tracked animals. As body feathers integrate the Hg  
 705 contamination over the year, we tested for a relationship with the  
 706 mean annual sea-ice values of tracked birds during the breeding  
 707 and non-breeding season.

## 708 Data Analyses

710 Data were analysed in R4.1.0., and visualized in R and in ArcGIS  
 711 10.2.2. Normality was tested using Shapiro tests and QQ plots.  
 712 Stable isotopes and Hg values were not normally distributed, and  
 713 univariate statistics were carried out using non-parametric tests,  
 714 while the data were successfully transformed using transform  
 715 Tukey in the R package “rcompanion” before carrying out  
 716 multivariate statistics such as linear models. A comparison of the  
 717 model outputs did not show any large differences between models  
 718 using transformed and untransformed data. Thus, effect plots are  
 719 given from models of untransformed data to enhance readability,  
 720 i.e., showing the actual scale of the data.

721 As Hg concentrations differed among the species and did  
 722 not show a linear relationship with stable isotope values, we  
 723 ran GAMs in the R package “mgcv”, separately for the species.  
 724 As proxies for the trophic position, we included either  $\delta^{15}\text{N}$  or  
 725 the estimated trophic position based on the linear regression of  
 726 feather  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  values ( $\text{TP}_{\text{LM}}$ ). As proxies for distribution,  
 727 we included either  $\delta^{13}\text{C}$  or the distribution zone. We checked  
 728 all GAMs for model convergence and random distribution of  
 729 residuals, and reported statistics (effective degrees of freedom and  
 730  $p$ -values) for the GAMs run separately for each parameter.

731 We further ran a model selection separately for  
 732 each species with the dredge function in the R  
 733 package MuMIn on the full models for feathers:  
 734  $\text{gam}(\text{THg.feathers} \sim \text{s}(\text{TP\_est}) + \text{s}(\delta^{13}\text{C.feathers}) +$   
 735  $\text{s}(\delta^{15}\text{N.feathers}) + \text{habitat} + \text{ocean})$ , and for blood:  
 736  $\text{gam}(\text{THg.blood} \sim \text{s}(\delta^{13}\text{C.blood}) + \text{s}(\delta^{15}\text{N.blood}) + \text{season} +$   
 737  $\text{habitat} + \text{ocean})$ . For the selected best models, we report the  
 738 coefficients and, as a measure of effect size, calculated eta  
 739 squared values ( $\eta^2$ ) obtained with the EtaSq function in the R  
 740

741 <sup>1</sup>movebank.org

742 package “DescTools”. Unless indicated otherwise, mean values  
 743 are given  $\pm$  SD.

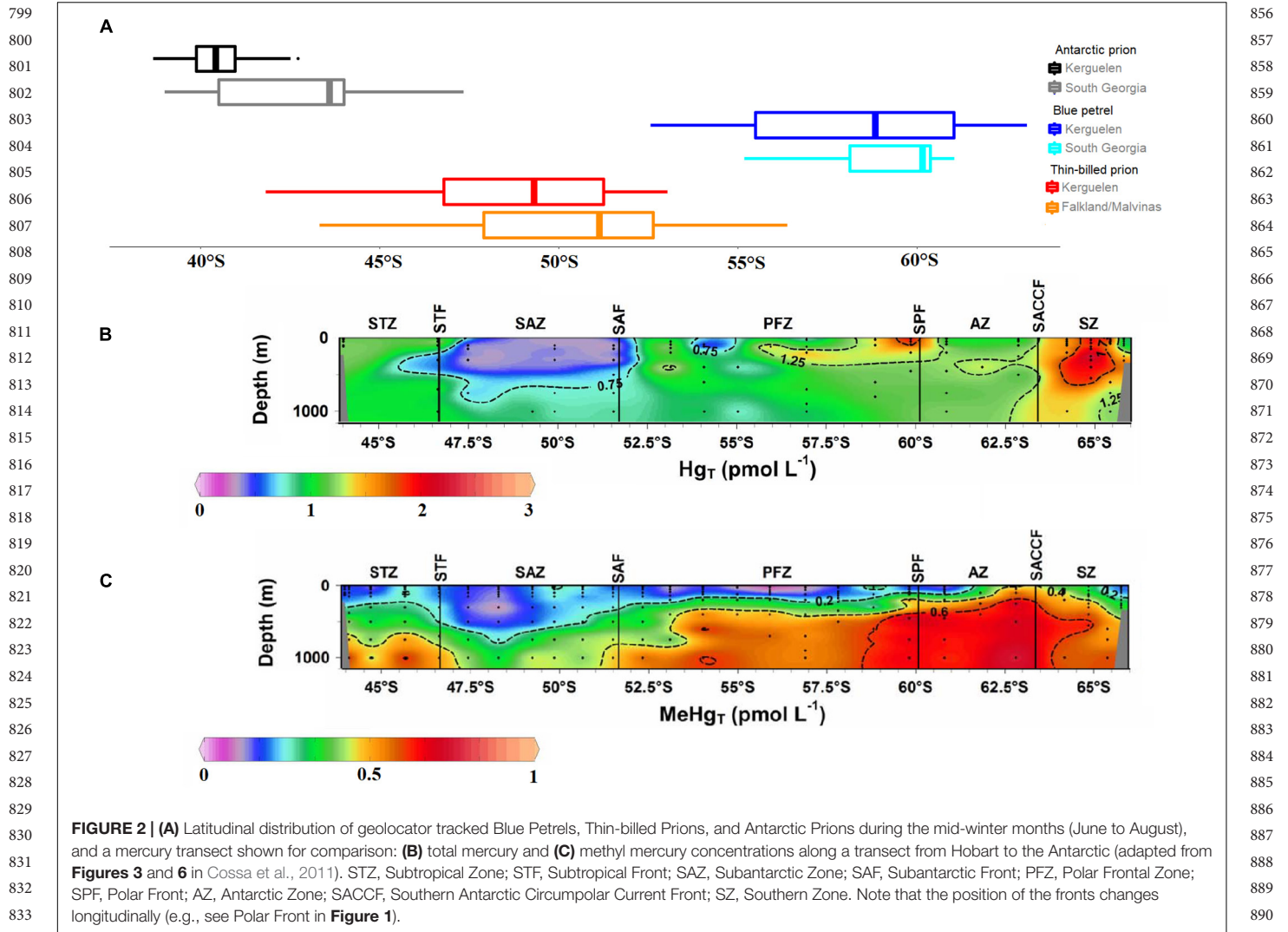
## 744 RESULTS

### 745 Year-Round Distribution and Moulting 746 Sites

747 The three species and their different populations had distinct  
 748 moulting sites and winter distributions (**Figures 1, 2**). Blue  
 749 Petrels and Thin-billed Prions moulted south of the Antarctic  
 750 Polar Front, and Antarctic Prions to its north. In all three species,  
 751 birds from Kerguelen started the core period of moult later  
 752 than birds from the south-west Atlantic colonies (mean 12 days,  
 753 11 days, and 28 days later in Blue Petrels, Thin-billed Prions,  
 754 and Antarctic Prions, respectively: **Supplementary Figure 1** and  
 755 **Supplementary Table 1**). Blue Petrels from South Georgia and  
 756 Kerguelen moulted in the Southern Ocean between 20°W and  
 757 30°E, overlapping between 20°W and 10°E (**Figure 1**). Based  
 758 on the immersion data, the core moult phase took place on  
 759 average between early February and late March in Blue Petrels  
 760 from South Georgia, and between mid-February and early April  
 761 in Blue Petrels from Kerguelen (**Supplementary Figure 1** and  
 762 **Supplementary Table 1**). The latitudes during the breeding  
 763 and moulting period differed only slightly for Blue Petrels  
 764 from Kerguelen and South Georgia (**Supplementary Figure 2**),  
 765 whereas ship-based observations indicate that Blue Petrels from  
 766 Diego Ramírez moult at higher latitudes (c. 70°S; Ryan et al.,  
 767 2020). Blue Petrels from Kerguelen and South Georgia spent the  
 768 mid-winter mostly south of 55°S (**Figure 2** and **Supplementary**  
 769 **Figure 2**). Although both populations moulted in the Atlantic  
 770 Ocean, subsequent longitudinal movements were in opposite  
 771 directions; birds from Kerguelen returned to the Indian Ocean,  
 772 whereas those from South Georgia entered the Pacific Ocean in  
 773 mid-winter (July–August) (**Supplementary Figure 2**).

774 The moulting areas of Thin-billed Prions were southeast  
 775 and southwest of the Falkland Islands, and most birds from  
 776 the Falklands and Kerguelen moulted in waters between 25°W  
 777 and 30°E, overlapping between 0° and 30°E (**Figure 1**). The  
 778 core moult period was between late February and early April  
 779 in Thin-billed Prions from the Falkland Islands, and between  
 780 early March and late April in Thin-billed Prions from Kerguelen  
 781 (**Supplementary Figure 1** and **Supplementary Table 1**). The  
 782 year-round latitudinal distribution was very similar for Thin-  
 783 billed Prions from both colonies (**Supplementary Figure 3**),  
 784 whereas longitudinal movements were in opposite directions  
 785 (**Supplementary Figure 3**). Thin-billed Prions spent the mid-  
 786 winter mostly between 45°S and 55°S, intermediate between the  
 787 other two species (**Figure 2** and **Supplementary Figure 3**).

788 Antarctic Prions generally moulted north of the Antarctic  
 789 Polar Front, and the moult areas of the birds from South Georgia  
 790 and Kerguelen did not overlap (**Figure 1**). The core moult took  
 791 place in the pre-breeding period, between late July and mid-  
 792 October in Antarctic Prions from South Georgia, and between  
 793 early August and late October in Antarctic Prions from Kerguelen  
 794 (**Supplementary Figure 1** and **Supplementary Table 1**). The  
 795 latitudes during the breeding period were slightly lower, and  
 796



those in the winter and moult periods slightly higher, for Antarctic Prions from Kerguelen (**Figure 2** and **Supplementary Figure 4**), and longitudinal movements were relatively short in this species (**Supplementary Figure 4**). Antarctic Prions spent the mid-winter mostly north of 45°S (**Figure 2** and **Supplementary Figure 4**), and moulted during this time. Antarctic Prions had longer core-moult periods (71 and 88 days) than the other two species (43–53 days, **Supplementary Table 1**).

### Feather Stable Isotope Values

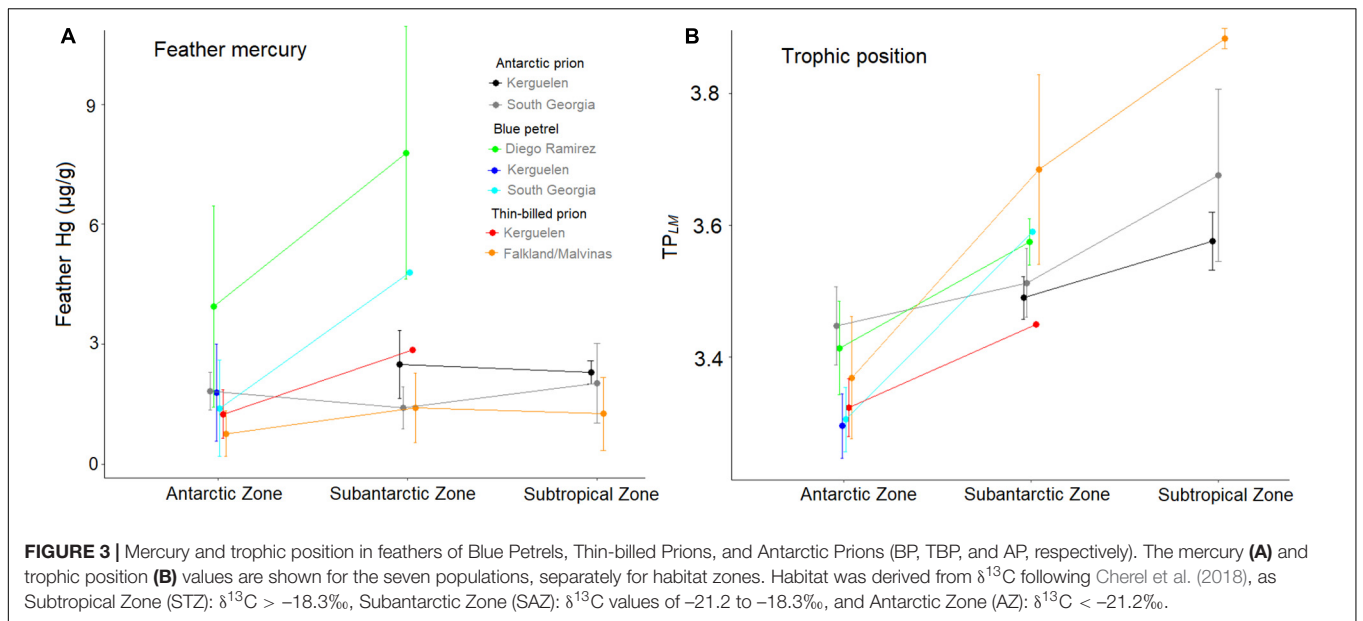
Stable isotope values of feathers differed among species (**Tables 1–3** and **Supplementary Figure 5**), with the  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values increasing from Blue Petrels to Thin-billed Prions to Antarctic Prions (Kruskal–Wallis tests; for  $\delta^{13}\text{C}$ :  $\chi^2 = 100.2$ ,  $d.f. = 2$ ,  $p < 0.001$ , *post-hoc* Dunn-tests: all  $p < 0.001$ , for  $\delta^{15}\text{N}$ :  $\chi^2 = 27.6$ ,  $d.f. = 2$ ,  $p < 0.001$ , *post-hoc* Dunn-tests: Blue Petrels vs. Thin-billed Prions  $p = 0.292$ , all other  $p < 0.001$ ). Trophic positions based on the subset of feathers analysed for CSIA from Thin-billed Prions and Blue Petrels ranged from 3.0 to 4.3. A linear model detected no

significant difference in trophic positions between the species (ANOVA tests;  $F_{1,26} = 1.04$ ,  $p = 0.316$ ,  $\eta^2 = 0.045$ ), whereas differences among the oceans were significant ( $F_{2,26} = 3.82$ ,  $p = 0.035$ ,  $\eta^2 = 0.227$ ), as were differences among feathers from AZ and SAZ distributions ( $F = 46.9$ ,  $p < 0.001$ ,  $\eta^2 = 0.511$ ; **Supplementary Figure 6**). Trophic positions were higher in the Pacific population, and birds with more northerly distributions (**Figure 3** and **Supplementary Figure 6**).

Across species, the trophic positions determined from feathers using linear models ( $TP_{LM}$ ), ranged from 3.2 to 3.9. Using this larger data set, we detected moderate differences in  $TP_{LM}$  among species ( $F_{2,206} = 109.4$ ,  $p < 0.001$ ,  $\eta^2 = 0.148$ ) and oceans ( $F_{2,206} = 23.2$ ,  $p < 0.001$ ,  $\eta^2 = 0.184$ ), and strong differences among distributions ( $F_{2,206} = 263.0$ ,  $p < 0.001$ ,  $\eta^2 = 0.574$ ). Trophic positions were elevated and highly variable in the Pacific population, and birds with more northerly distributions (**Figure 4**).

Mercury concentrations in feathers differed among species (Kruskal–Wallis ANOVA:  $\chi^2 = 85.5$ ,  $d.f. = 2$ ,  $p < 0.001$ , *post-hoc* Dunn-tests: Blue Petrels vs. Antarctic Prions  $p = 0.265$ , all





other  $p < 0.001$ ). The highest mean Hg concentrations were in Blue Petrels ( $2.17 \pm 1.94 \mu\text{g/g}$ ), followed by Antarctic Prions ( $1.85 \pm 0.75 \mu\text{g/g}$ ), and Thin-billed Prions ( $1.14 \pm 0.69 \mu\text{g/g}$ ). Of the seven populations, Blue Petrels from Diego Ramírez (Pacific Ocean) had much higher Hg concentrations than predicted by their latitudinal distribution and trophic positions (Figure 4 and Supplementary Figures 7–9).

Generalized Additive Models (Figure 4 and Table 4) showed a significant effect of ocean basin in all three species, with the most elevated Hg values in the Pacific Ocean and the lowest in the Atlantic Ocean (Figure 5). In Blue Petrels and Thin-billed Prions, distribution ( $\delta^{13}\text{C}$ , habitat zone) as well as trophic position ( $\delta^{15}\text{N}$ ,  $\text{TP}_{LM}$ ) influenced Hg values (Figure 4 and Table 4). Model selection retained only ocean basin for Antarctic Prions (Figure 6), but all parameters except habitat zone for Blue Petrels and Thin-billed Prions. Coefficients for the effect of trophic position ( $\delta^{15}\text{N}$ ,  $\text{TP}_{LM}$ ) on feather Hg indicated a strong positive relationship for Blue Petrels, a weaker, negative relationship for Thin-billed Prions, and no influence for Antarctic Prions (Figure 6).

### Blood Stable Isotope Values

Mean blood  $\delta^{13}\text{C}$  values were lowest in Blue Petrels ( $-24.4 \pm 1.4\text{‰}$ ), and higher in Thin-billed Prions ( $-21.4 \pm 2.6\text{‰}$ ) and Antarctic Prions ( $-22.5 \pm 1.3\text{‰}$ , Kruskal–Wallis ANOVA:  $\chi^2 = 90.3$ ,  $d.f. = 2$ ,  $p < 0.001$ ), with no significant difference between the last two species (*post-hoc* Dunn-tests: Thin-billed vs. Antarctic Prions  $p = 0.474$ , all other  $p < 0.001$ ). Blood  $\delta^{15}\text{N}$  values differed among species, and were lowest in Antarctic Prions ( $8.3 \pm 0.3\text{‰}$ ), intermediate in Blue Petrels ( $9.0 \pm 0.8\text{‰}$ ), and highest in Thin-billed Prions ( $10.1 \pm 1.9\text{‰}$ , Kruskal–Wallis ANOVA:  $\chi^2 = 38.5$ ,  $d.f. = 2$ ,  $p < 0.001$ , *post-hoc* Dunn-tests: all  $p < 0.001$ ).

The trophic positions based on the subset of blood samples analysed for CSIA ranged from 3.3 to 4.0 in Thin-billed Prions

( $3.5 \pm 0.1$ ) and Blue Petrels ( $3.5 \pm 0.2$ ). According to  $\text{TP}_{CSIA}$  values, the trophic positions of the two species did not differ significantly (*t*-test,  $t = -0.7$ ,  $d.f. = 11.6$ ,  $p = 0.480$ ).

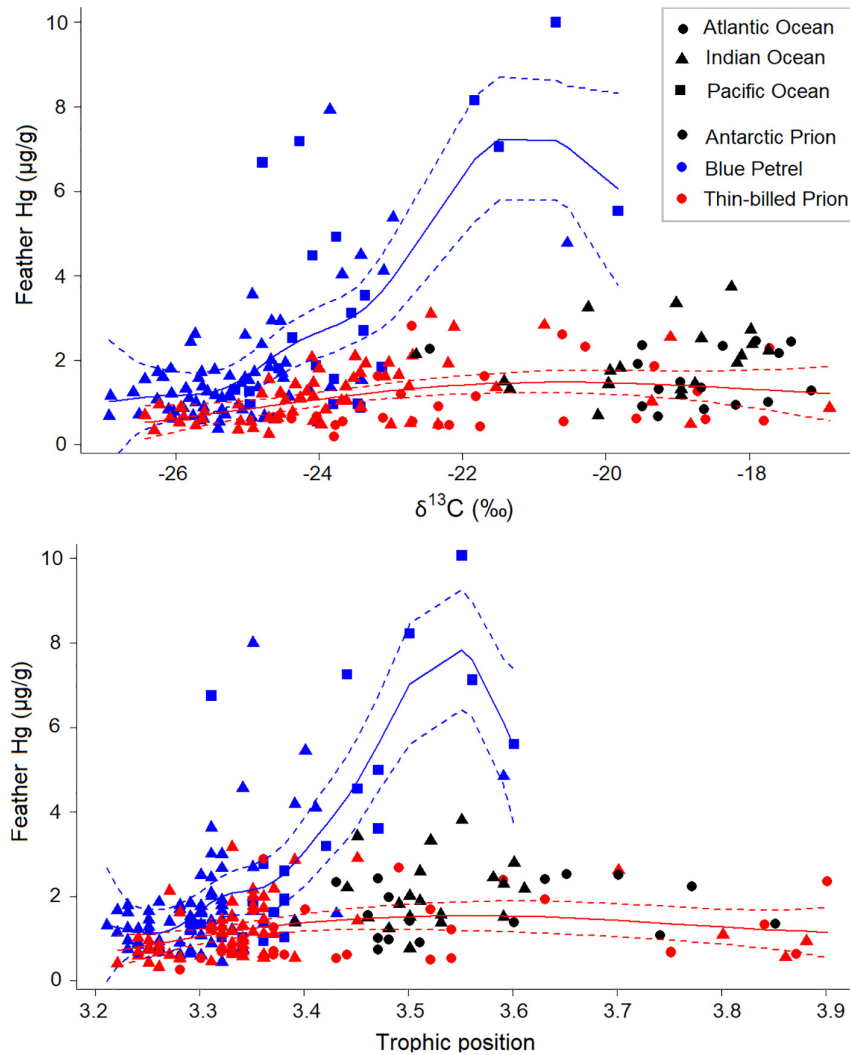
Mean Hg concentrations in blood differed among species (Kruskal–Wallis ANOVA:  $\chi^2 = 124.0$ ,  $d.f. = 2$ ,  $p < 0.001$ , *post-hoc* Dunn-tests: all  $p < 0.001$ ), with the highest concentrations in Blue Petrels ( $2.99 \pm 1.97 \mu\text{g/g}$ ), then Thin-billed Prions ( $0.99 \pm 0.41 \mu\text{g/g}$ ), and Antarctic Prions ( $0.51 \pm 0.22 \mu\text{g/g}$ ).

Species-specific GAMs showed a significant effect of latitudinal distribution ( $\delta^{13}\text{C}$ , habitat zone) in all three species (Table 5 and Figures 7, 8). However, this was only clearly positive in Blue Petrels (Figures 7, 8 and Supplementary Figure 10). The trophic position ( $\delta^{15}\text{N}$ ) influenced Hg values in Blue Petrels and Thin-billed Prions (Table 5), with a clear increase only in Blue Petrels (Figure 8). There was a significant effect of ocean basin for Antarctic and Thin-billed Prions (Table 5). Changes in Hg and stable isotope values over the season were apparent in blood of Blue Petrels and, to a lesser extent, of Thin-billed Prions (Table 5 and Figure 7). There was a decrease of an order of magnitude in Hg concentrations in blood of Blue Petrels, which were sampled from arrival in September to the post-moult visit to the colony in April (Figure 9).

For blood Hg, all parameters except  $\delta^{15}\text{N}$  and habitat were retained in the best models for Antarctic Prions (Figure 6D). Habitat was also excluded for Blue Petrels (Figure 6E), and in three of four best models for Thin-billed Prions (Figure 6F). Coefficients for the effect of trophic position ( $\delta^{15}\text{N}$ ) on the feather Hg indicated a strong positive relationship for Blue Petrels, but values close to zero for Thin-billed Prions and Antarctic Prions (Figure 6).

### Sea-Ice Concentration

The year-round sea-ice concentration in areas used by tracked Blue Petrels, Thin-billed Prions and Antarctic Prions (Figure 10) from the Atlantic and Indian Ocean showed two annual peaks:



**FIGURE 4 |** Species-specific Generalized Additive Model (GAM) model fits for mercury values in feathers of Antarctic Prions, Blue Petrels, and Thin-billed Prions. Estimated smoothing curves for mercury values in feathers in relation to  $\delta^{13}\text{C}$  and trophic position (derived from  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  values, with 95% confidence intervals) are given where statistically significant. For GAM statistics, see **Table 4**.

in April for Blue Petrels and Thin-billed Prions from Atlantic colonies, and again in August–September for Blue Petrels. Blue Petrels from the Indian Ocean had higher sea-ice overlap than birds from the Atlantic in April to August (**Figure 10**). Blue Petrels from Diego Ramírez have not yet been tracked (but see distribution in **Figure 1**). The highest exposure to sea ice was in September for all populations except Thin-billed Prions from New Island (Falklands) (**Figure 10**).

During the period of wing moult (**Supplementary Figure 1** and **Supplementary Table 1**), sea-ice exposure was low ( $<0.01$ ) for all populations.

### Data From Individually Tracked Birds

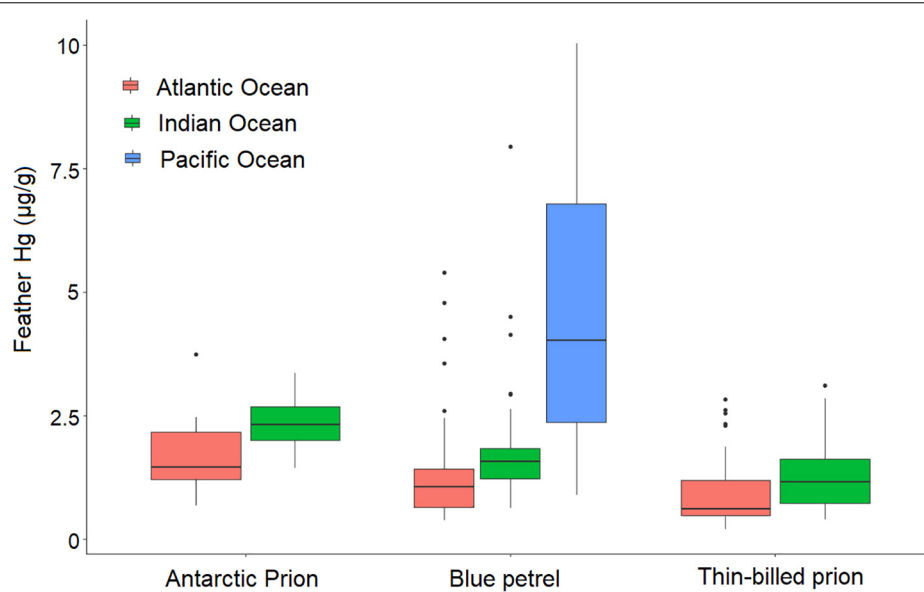
Matching data on blood Hg and sea-ice exposure were available for tracked individuals from four populations (**Figure 11A**), and on feather Hg and sea-ice exposure for five populations

(**Figure 11B**). Model selection suggested that species differences were sufficient to explain differences in blood Hg values, but when analysing the dataset across species, a GAM suggested that blood Hg values increased with maximum sea-ice exposure (**Figure 11A** and **Table 6**). In contrast, mean annual sea-ice exposure was not related to feather Hg concentrations (**Figure 11B** and **Table 6**).

## DISCUSSION

In the present study, we examined temporal and spatial effects on stable isotope values and Hg concentrations in seven populations of three species of small petrels in widely separated oceans. We found evidence that higher trophic level and the distribution may result in higher exposure to Hg. We also found a carry-over effect of Hg exposure between wintering and breeding grounds.





**FIGURE 5 |** Mercury values in feathers of Antarctic Prions, Blue Petrels, and Thin-billed Prions from different ocean basins. Boxplots showing medians, interquartile ranges, and outliers.

**TABLE 4 |** Generalized Additive Model (GAM) results for feather mercury values, separately for the species, as a function of distribution ( $\delta^{13}\text{C}$ , habitat), trophic position ( $\delta^{15}\text{N}$ ,  $\text{TP}_{LM}$ ), and ocean basin (Atlantic, Indian, or Pacific).

Species	Variable	Smoother edf (P)	Effect size	Estimate (SE) P
Blue Petrel (n = 90)	$\delta^{13}\text{C}$	<b>5.29 (P &lt; 0.001)</b>	<b>0.532</b>	
	$\delta^{15}\text{N}$	<b>5.58 (P &lt; 0.001)</b>	<b>0.536</b>	
	$\text{TP}_{LM}$	<b>6.09 (P &lt; 0.001)</b>	<b>0.592</b>	
	Habitat		<b>0.198</b>	<b>4.78 (1.02)</b> <b>P &lt; 0.001</b>
	Ocean		<b>0.299</b>	<b>2.90 (0.51)</b> <b>P &lt; 0.001</b>
Thin-billed Prion (n = 87)	$\delta^{13}\text{C}$	<b>2.23 (P = 0.002)</b>	<b>0.176</b>	
	$\delta^{15}\text{N}$	1.57 (P = 0.523)	0.023	
	$\text{TP}_{LM}$	<b>2.35 (P = 0.029)</b>	<b>0.120</b>	
	Habitat		<b>0.049</b>	<b>0.46 (0.22)</b> <b>P = 0.043</b>
	Ocean		<b>0.050</b>	<b>0.31 (0.15)</b> <b>P = 0.037</b>
Antarctic Prion (n = 36)	$\delta^{13}\text{C}$	1.30 (P = 0.699)	0.030	
	$\delta^{15}\text{N}$	1.70 (P = 0.454)	0.073	
	$\text{TP}_{LM}$	1.62 (P = 0.562)	0.056	
	Habitat		0.077	0.13 (0.41) P = 0.267
	Ocean		<b>0.205</b>	<b>0.75 (0.25)</b> <b>P = 0.006</b>

Habitat was derived from  $\delta^{13}\text{C}$  following Cherel et al. (2018), as Subtropical Zone (STZ):  $\delta^{13}\text{C} > -18.3\text{‰}$ , Subantarctic Zone (SAZ):  $\delta^{13}\text{C}$  values of  $-21.2$  to  $-18.3\text{‰}$ , and Antarctic Zone (AZ):  $\delta^{13}\text{C} < -21.2\text{‰}$ . GAM results are reported for separate models for each parameter. Parameters with a statistically significant effect on feather mercury values are marked bold.

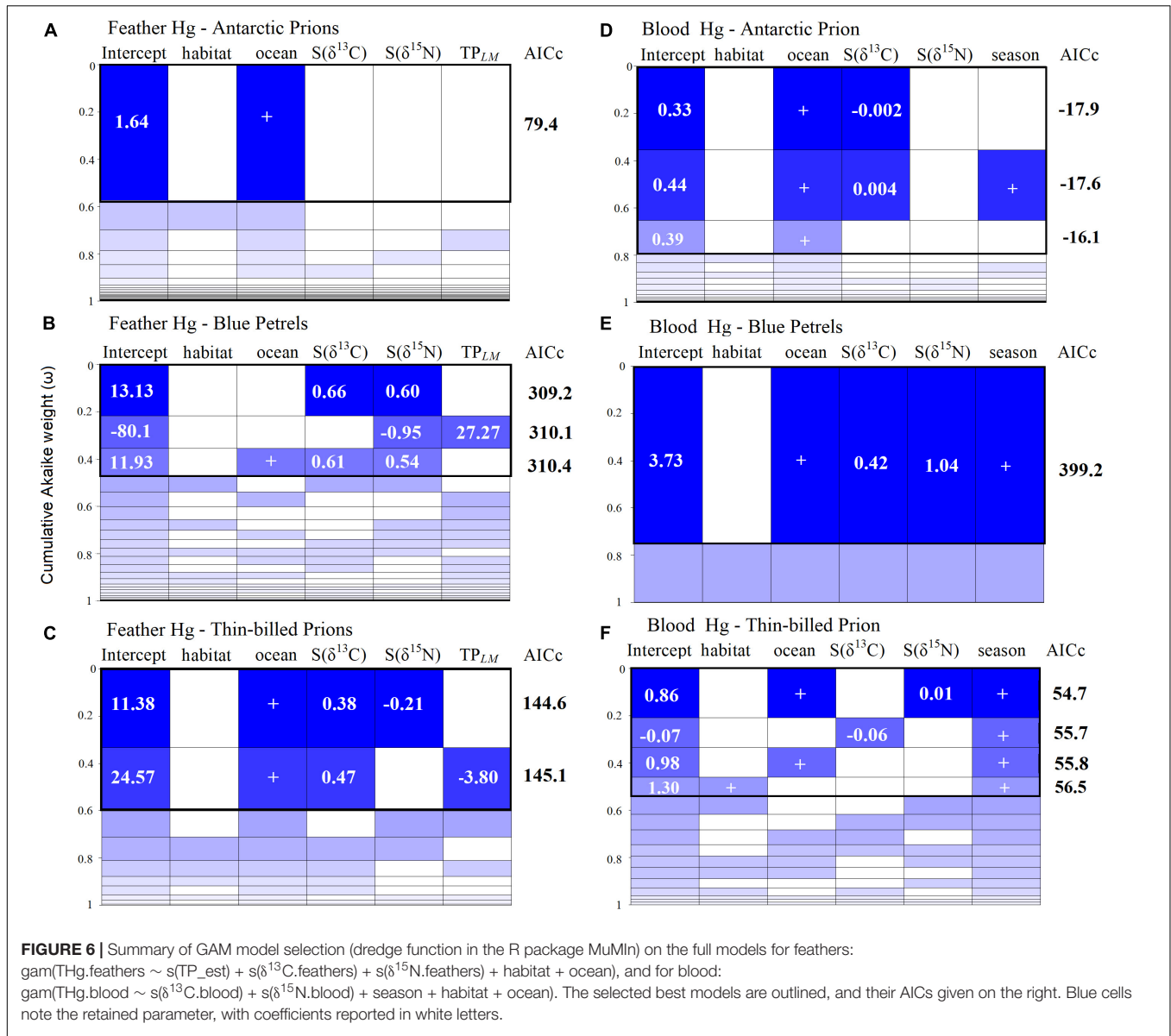
## Variation Among Species and Populations in Mercury Concentrations

We found interspecific differences in Hg concentrations in both blood and feathers, with the highest value for both tissues in Blue Petrels. In the literature, differences among

species in Hg concentrations are mostly discussed in relation to biomagnification processes and thus, trophic position (e.g., Becker et al., 2002; Anderson et al., 2009; Blévin et al., 2013; Gatt et al., 2020). However, we here compared three small-bodied petrel species of similar trophic positions, according to  $\delta^{15}\text{N}$

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values in feathers and blood samples. We found that similar trophic position in different water masses did not lead to the same degree of Hg biomagnification. For example, Thin-billed Prions had the highest trophic positions relative to their distribution, but lower Hg concentrations than Blue Petrels. This result does not agree with the suggestion that the trophic position is the most important factor explaining variation in Hg concentrations in Southern Ocean seabirds (Becker et al., 2002). Likewise, in tunas trophic effects (i.e., geographical changes in foraging ecology) had a limited influence on the spatial variability of tissue Hg concentrations (Médieu et al., 2022).

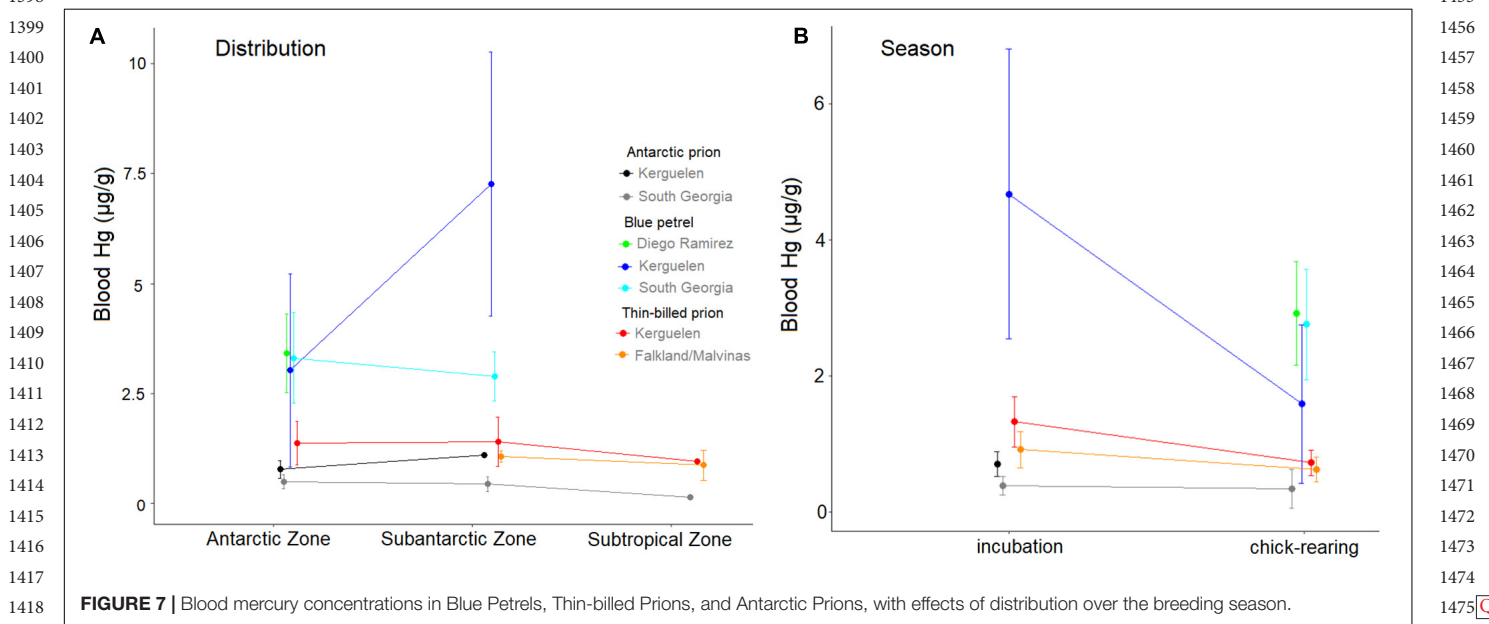
Despite generally low trophic positions, dietary differences exist among the species, especially in the relative importance of fish. At South Georgia, crustaceans, and particularly Antarctic krill (*Euphausia superba*), predominated in Antarctic Prion and Blue Petrel diets, but fish was considerably more important

for the Blue Petrels (Prince, 1980). In Blue Petrels at Marion Island (Steele and Klages, 1986), the diet consisted of 60% crustaceans, 21% myctophid fish and 16% squid by mass. In Blue Petrels at Kerguelen, however, the contribution of fish was higher (57%, Cherel et al., 2002b). Compared to King Penguins (*Aptenodytes patagonicus*) at Kerguelen which have a diet consisting of primarily (>90%) myctophids, Blue Petrels at the same island group have only slightly lower feather Hg values (Table 1, King Penguins = 2.2 ± 0.5 μg/g; Carravieri et al., 2013). In comparison, the proportion of fish taken by Thin-billed Prions and Antarctic Prions is very low both in Kerguelen (Cherel et al., 2002a) and the Falkland Islands (Quillfeldt et al., 2010). The hyperiid amphipod *Themisto gaudichaudii* was consistently the dominant prey item for Thin-billed prions. These predatory pelagic crustaceans may be responsible for the relatively high trophic position of Thin-billed Prions (Figure 3), but result in

1369 **TABLE 5** | Generalized Additive Model (GAM) results for blood mercury values, separately for the species, as a function of distribution ( $\delta^{13}\text{C}$ , habitat), trophic position ( $\delta^{15}\text{N}$ ), period (early = arrival to incubation vs. late = chick-rearing), and ocean basin (Atlantic, Indian, or Pacific). 1426

Species	Variable	Smoother edf ( <i>P</i> )	Effect size	Estimate (SE) <i>P</i>
Blue Petrel ( <i>n</i> = 135)	$\delta^{13}\text{C}$	<b>2.99 (<i>P</i> &lt; 0.001)</b>	<b>0.498</b>	
	$\delta^{15}\text{N}$	<b>1.76 (<i>P</i> &lt; 0.001)</b>	<b>0.546</b>	
	Period		<b>0.373</b>	<b>-2.54 (0.28)</b> <b><i>P</i> &lt; 0.001</b>
	Ocean		0.002	0.28 (0.53) <i>P</i> = 0.602
	Habitat		<b>0.196</b>	<b>2.76 (0.48)</b> <b><i>P</i> &lt; 0.001</b>
Thin-billed Prion ( <i>n</i> = 120)	$\delta^{13}\text{C}$	<b>2.85 (<i>P</i> &lt; 0.001)</b>	<b>0.321</b>	
	$\delta^{15}\text{N}$	<b>4.70 (<i>P</i> &lt; 0.001)</b>	<b>0.356</b>	
	Period		<b>0.339</b>	<b>-0.50 (0.06)</b> <b><i>P</i> &lt; 0.001</b>
	Ocean		<b>0.238</b>	<b>0.40 (0.07)</b> <b><i>P</i> &lt; 0.001</b>
	Habitat		<b>0.238</b>	<b>-0.42 (0.07)</b> <b><i>P</i> &lt; 0.001</b>
Antarctic Prion ( <i>n</i> = 26)	$\delta^{13}\text{C}$	<b>1.00 (<i>P</i> = 0.002)</b>	<b>0.344</b>	
	$\delta^{15}\text{N}$	1.33 ( <i>P</i> = 0.353)	0.087	
	Period		0.050	-0.19 (0.16) <i>P</i> = 0.271
	Ocean		<b>0.503</b>	<b>0.32 (0.06)</b> <b><i>P</i> &lt; 0.001</b>
	Habitat		<b>0.311</b>	<b>-0.20 (0.08)</b> <b><i>P</i> = 0.014</b>

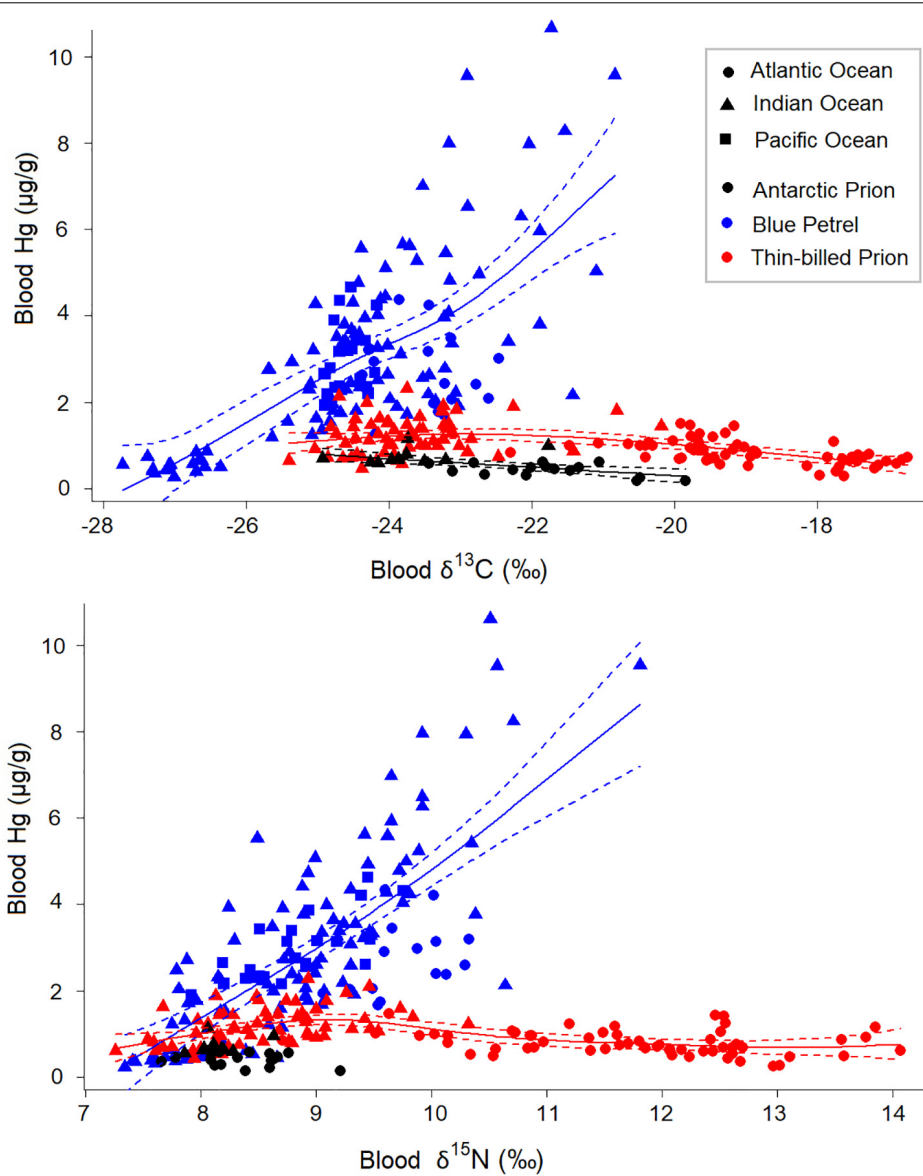
1395 *Habitat* was derived from  $\delta^{13}\text{C}$  following Jaeger et al. (2010), as Subtropical Zone (STZ):  $\delta^{13}\text{C}$  > -20.1‰, Subantarctic Zone (SAZ):  $\delta^{13}\text{C}$  values of -22.9 to -20.1‰, 1452  
 1396 and Antarctic Zone (AZ):  $\delta^{13}\text{C}$  < 22.9‰. GAM results are reported for separate models for each parameter, and parameters with a statistically significant effect on blood 1453  
 1397 mercury values are marked bold. 1454  
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1421 little Hg take-up. At Kerguelen, Hg concentrations were higher 1478  
 1422 in myctophid fish (up to 0.424 µg/g dw) and, to a lesser extent, 1479  
 1423 squid (up to 0.270 µg/g dw) compared to crustaceans (up to 1480  
 1424 0.034 in amphipods, 0.074 in copepods and 0.125 in euphasiids) 1481  
 1425 (Cipro et al., 2018), and fish in the diet was suggested to be 1482

the most important driver of elevated Hg values in seabirds 1478  
 (Bocher et al., 2003). 1479

While Blue Petrels are the most piscivorous of the species in 1480  
 the present study, they also use the most southerly habitats over 1481  
 the non-breeding season (Quillfeldt et al., 2015; **Figure 2**). Blue 1482

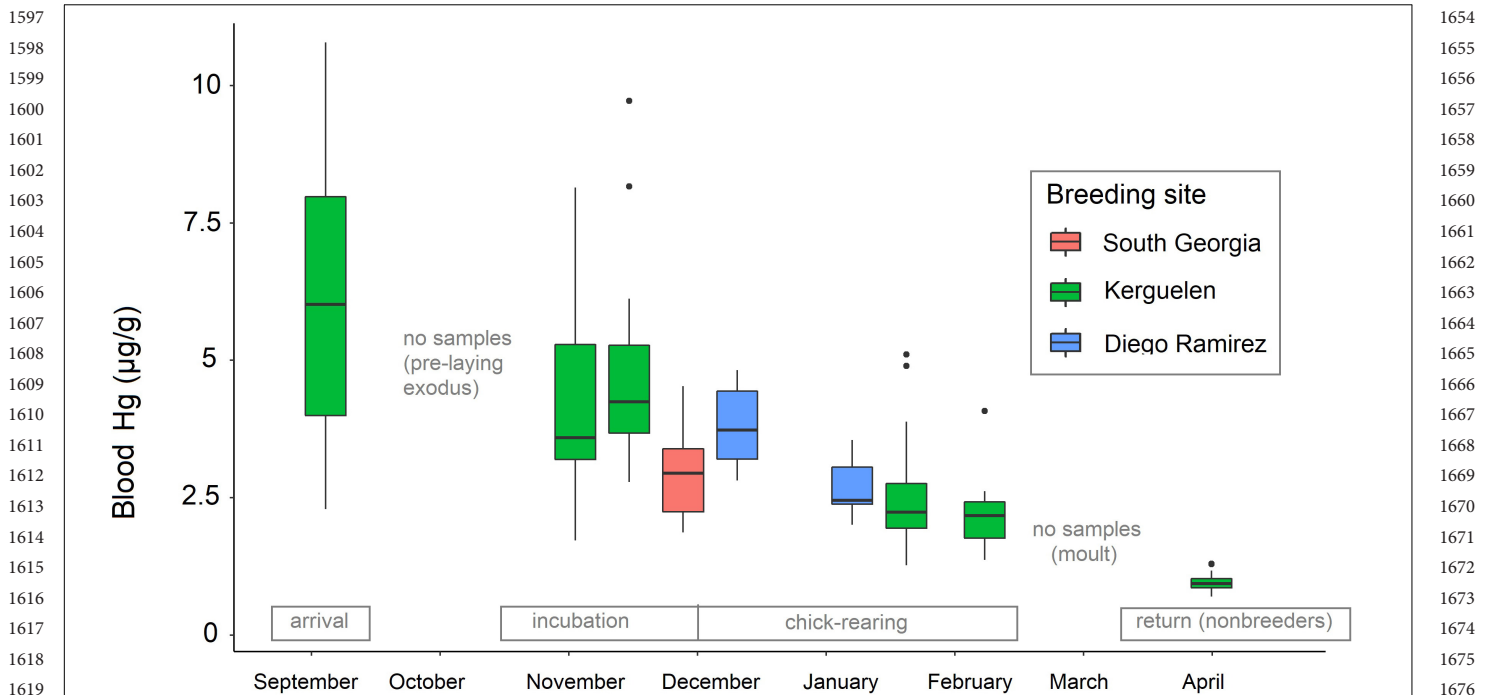


**FIGURE 8 |** Species-specific Generalized Additive Model (GAM) model fits for mercury values in blood of Antarctic Prions, Blue Petrels, and Thin-billed Prions. Estimated smoothing curves for mercury values in blood in relation to  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  are given with 95% confidence intervals where statistically significant. For GAM statistics, see **Table 5**.

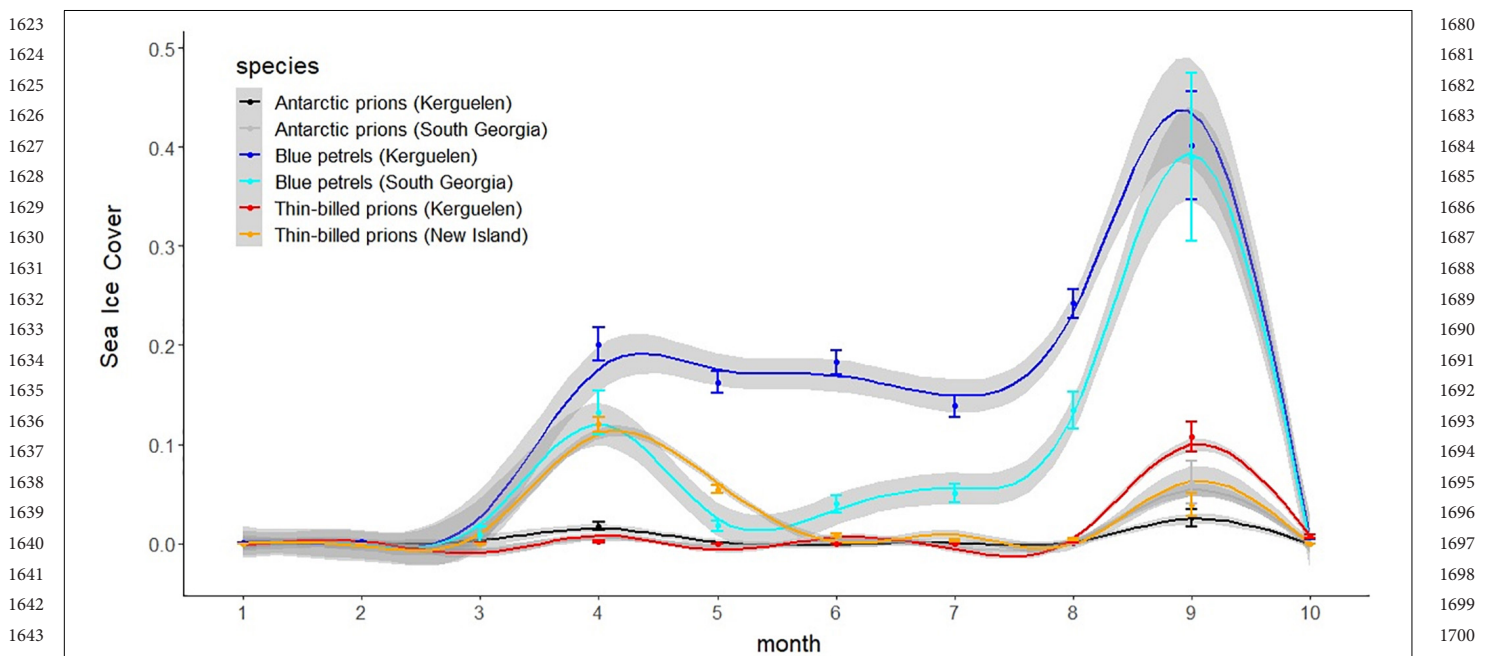
Petrels from Kerguelen spent the winter in waters with >10% sea-ice (Figure 10), and all Blue Petrels spent time in waters with 30–40% sea-ice before the start of the breeding season in August–September (Figure 10). Observations off west Antarctica suggested that Blue Petrels avoided areas with dense pack ice, but were found just outside the marginal ice zone, at sea surface temperatures of  $-0.7$  to  $0.9^\circ\text{C}$  (Ryan et al., 2020). Mercury measurements along a transect from Hobart to the Antarctic (Cossa et al., 2011; see Figure 2) identified two zones of elevated dissolved Hg concentrations: in the Southern Zone where it is caused by processes in the ice-atmosphere-ocean interface like brine formation, and south of the Antarctic Polar Front (Cossa et al., 2011). In the Southern Zone, there is further a build-up

of MeHg-enriched surface waters during winter months, when the sea-ice extent increases and the sea surface is protected from the UV and, thus, from MeHg photo-reduction (Cossa et al., 2011). However, the MeHg concentration was highest close to the Southern Antarctic Circumpolar Current Front, due to upwelling of waters from the minimum oxygen zone (Cossa et al., 2011; see Figure 2).

Some Antarctic seabirds have a strong affinity to the sea-ice environment, in particular Snow Petrels (*Pagodroma nivea*), Antarctic Petrels (*Thalassoica antarctica*), Adélie Penguins (*Pygoscelis adeliae*), and Emperor Penguins (*Aptenodytes forsteri*). As Procellariiformes (albatrosses, shearwaters, petrels, and storm-petrels) tend to have higher feather Hg concentrations



**FIGURE 9 |** Temporal changes in mercury concentrations in blood of Blue Petrels from different breeding locations over the breeding season.

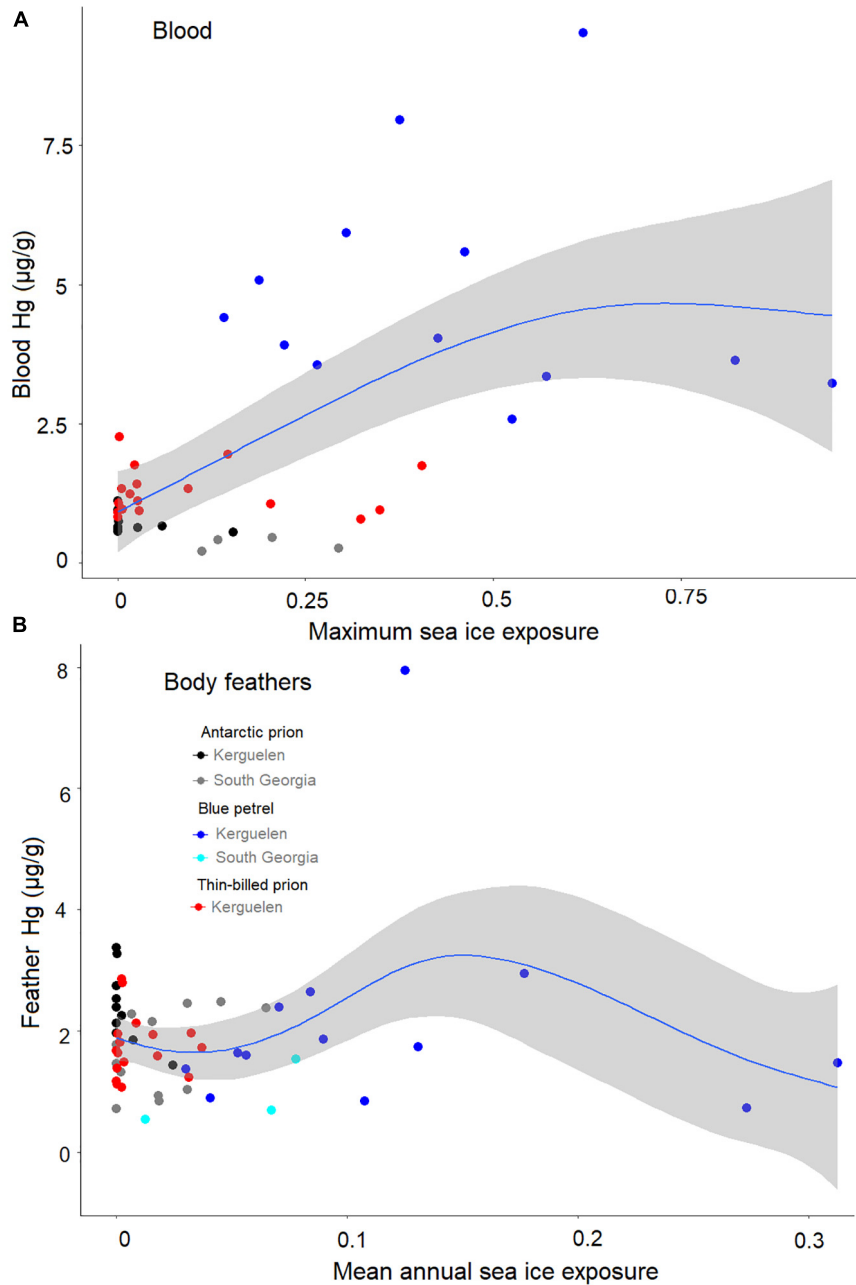


**FIGURE 10 |** Year-round percentage of sea-ice concentration (mean ± SD, and smooth lines with 95% confidence intervals) of geolocator tracked Blue Petrels, Thin-billed Prions, and Antarctic Prions.

than other species owing to their protracted moulting periods (Braune and Gaskin, 1987; Stewart et al., 1999), their Hg concentrations are particularly relevant here. However, a comparison with these species shows no particularly elevated Hg concentrations. In Snow Petrels from Adélie Land, the blood

Hg concentration averaged  $2.7 \pm 1.1$  (range: 1.0–5.3) µg/g dw in the pre-laying season (Tartu et al., 2014), lower than the values in our study for Blue Petrels in the early breeding season (Figure 9). Likewise, Antarctic Petrels had moderate mean Hg concentrations in feathers ( $2.41 \pm 0.83$  µg/g dw) and blood





**FIGURE 11 |** Mercury concentrations in relation to the year-round sea-ice exposure of tracked Blue Petrels, Thin-billed Prions, and Antarctic Prions, shown for blood samples (A) and body feathers (B). Grey shaded areas show GAM smoothed 95% confidence intervals, obtained in the R package “ggplot2”, across the whole dataset.

cells ( $1.38 \pm 0.43 \mu\text{g/g dw}$ ; Carravieri et al., 2021). Similarly, Adélie Penguins and Emperor Penguins from the Ross Sea had low feather Hg concentrations ( $0.592 \pm 0.015 \mu\text{g/g}$  and  $1.351 \pm 0.058 \mu\text{g/g}$ , respectively; Pilcher et al., 2020). Therefore, the high values observed for Blue Petrels are unlikely to be explained directly by foraging in southern waters with up to 40% sea-ice concentration, but might have a connection with fish that migrate to the surface from the oxygen minimum layer, and with the elevated MeHg concentration close to the Southern

Antarctic Circumpolar Current Front and, thus, in waters from the minimum oxygen zone (Cossa et al., 2011; see Figure 2). Further research should be dedicated to test this hypothesis.

In the Arctic, Ivory Gulls (*Pagophila eburnean*) have the highest Hg concentrations in their eggs of any Arctic bird (Miljeteig et al., 2009; Bond et al., 2015). They consume ice-associated marine fish and scavenge on marine mammal carcasses. While the trophic position remained unchanged between 1877 and 2007 in ivory gulls from Arctic Canada

**TABLE 6** | Generalized Additive Model (GAM) results for tracked individuals, separately for blood and feathers, as a function of species and sea ice cover.

Tissue	Variable	Smoother edf (P)	Effect size	Estimate (SE) P
Blood	Species		<b>0.729</b>	<b>4.22 (0.43)</b> <b>P &lt; 0.001</b>
	Sea-ice cover (max)	<b>2.24 (P &lt; 0.001)</b>	<b>0.395</b>	
Feathers	Species		0.321	-0.25 (0.35) P = 0.478
	Sea-ice cover (mean)	3.98 (P = 0.072)	0.209	

GAM results are reported for separate models for each parameter, and parameters with a statistically significant effect on mercury values are marked bold.

and western Greenland (Bond et al., 2015), their feather Hg concentration increased by a factor of 45 (from 0.09 to 4.11  $\mu\text{g/g}$ ). Due to human activities such as coal and oil combustion, cement production, waste incineration, mining, smelting, and other industrial processes, the total and bioavailable amounts of Hg have dramatically increased in the environment since the industrial revolution (Pirrone et al., 2010; Arctic Monitoring and Assessment Programme [AMAP], 2019). The concentration of Hg that causes deleterious effects in birds depends on different factors, including diet composition, moult duration and the ability to demethylate Hg in the liver (Heinz et al., 2009), and has been given as 5–40  $\mu\text{g/g}$  in feathers in general (Burger and Gochfeld, 1997), or 10–15  $\mu\text{g/g}$  in piscivorous divers (Evers et al., 2014). All values observed here were below 10  $\mu\text{g/g}$ , but the highest values in Blue Petrels approached this concentration (Figure 4), warranting further monitoring in the future.

## Temporal Differences in Mercury Concentrations

We found temporal differences in Hg concentrations in blood samples, which were most pronounced in Blue Petrels. The highest concentrations were noted in September, after arrival from the wintering grounds, indicating that the adults arrived from Hg contaminated water masses or after feeding on prey with high Hg levels, but then switched to less contaminated prey. Hg in blood then decreased continually over the breeding season, and reached very low levels in birds sampled after returning to the colony post-moult. This was paralleled by a decline in trophic position, as indicated by  $\delta^{15}\text{N}$  values. Results of a previous study showed that mean  $\delta^{15}\text{N}$  values in adult Blue Petrels at Kerguelen decreased continuously throughout the annual cycle, from  $9.5 \pm 1.1\text{‰}$  on arrival in the colony in September, to  $7.3 \pm 0.5\text{‰}$  in the immediate post-breeding period in April to May (Cherel et al., 2014). In the present study, we measured a similar decrease from means of  $10.3 \pm 0.8\text{‰}$  on arrival in the colony in September to  $7.9 \pm 0.4\text{‰}$  at the post-nuptial stage in April (Table 1). The very low levels in April can be directly related to the Hg reset after Hg depuration in feathers, at a time when  $\delta^{15}\text{N}$  values are also very low (most likely indicating feeding on Antarctic krill).

Antarctic Prions did not show a pronounced seasonal trend in Hg concentrations, but had low values throughout the

breeding season. In both Thin-billed Prion populations and all years, the blood Hg values were somewhat higher early in the breeding season. Diet composition of Thin-billed Prions at the Falkland Islands changes during the breeding season: from 60% squid and 35% amphipods during incubation to more equal proportions of amphipods, krill and squid during chick rearing (Quillfeldt et al., 2010).

Mercury in blood represents two components: Hg incorporated from the diet during blood formation, and Hg stored in other tissues, such as the liver, kidney and muscles, since the last feather moult. Residual Hg in other tissues is thought to equilibrate with levels in muscle (especially MeHg; Renedo et al., 2021) and liver, which act as the main storage organs for Hg between moults (Bearhop et al., 2000b). It has been shown that a carry-over of Hg can occur from remote places, such that high exposure in winter may lead to elevated blood Hg values until late in the summer (Lavoie et al., 2014). Especially for individuals with high winter exposure to Hg, slow changes in blood Hg over time were reported, suggesting a fast uptake rate and slow depuration (Lavoie et al., 2014). Carry-over of Hg among seasons would also explain the temporal patterns observed in Blue Petrels in our study.

## Spatial Differences in Mercury

In the Blue Petrels, the population of Diego Ramírez most likely moulted off west Antarctica, i.e., in the Pacific Ocean sector of the Southern Ocean, from 67 to 71°S and 78 to 119°W (Ryan et al., 2020). Ryan et al. (2020) observed large numbers of moulting Blue Petrels sitting on the water in dense flocks in mid-February, which is in line with the  $10.7 \pm 2.5$  h per day spent sitting on the water by Blue Petrels from Kerguelen during moult (Cherel et al., 2016). Ryan et al. (2020) suggested that most of the birds observed in west Antarctica probably breed at Diego Ramírez, and this is also suggested by a comparison with distribution data of Blue Petrels from other colonies. Blue Petrels from both Kerguelen and South Georgia were found in the Atlantic sector of the Southern Ocean (20°W to 30°E) in March, during the core moulting period, and thus far away from the moulting aggregations observed off west Antarctica (Ryan et al., 2020).

Latitudinal differences in distribution influence Hg exposure, with lower Hg in Antarctic waters compared to subantarctic waters, a trend reported in previous studies (e.g., Carravieri et al., 2014, 2016, 2017, 2020; Cherel et al., 2018). We found no further increase towards subtropical waters. The trophic position also increased from polar to subantarctic waters, but continued to increase to subtropical waters. Thus, differences in trophic position would not fully explain the observed patterns. Indeed, a more detailed analysis revealed that not all populations show an increase in blood Hg concentrations associated with  $\delta^{13}\text{C}$  values. Differences in prey as well as carry-over effects of Hg may mask spatial differences.

Further spatial differences in Hg values were observed when comparing populations from different ocean sectors. Values were lowest in the Atlantic Ocean, intermediate in the Indian Ocean, and highest in the Pacific Ocean, although this was only based on one population. That population, Blue Petrels from Diego Ramírez, had high feather Hg concentrations ( $4.42 \pm 2.72$   $\mu\text{g/g}$ )

dw), comparable with Blue Petrels on Marion Island, Indian Ocean ( $4.62 \pm 4.11 \mu\text{g/g dw}$ , **Supplementary Table 2**). Tracking and dietary data are still lacking from both populations.

A difference in Hg has also been observed for other organisms such as Marbled Rockcod (*Notothenia rossii*), where mean muscle Hg concentrations of fish in waters around Kerguelen ( $0.255 \mu\text{g/g dw}$ ; Bustamante et al., 2003) were three times higher than in the South Shetland Islands in the Atlantic sector of the Southern Ocean ( $0.077 \mu\text{g/g dw}$ ; Cipro et al., 2017). Such differences may be due to differences in Hg sources and oceanographic features.

## CONCLUSION

In line with previous studies, we found high Hg concentrations in Blue Petrels. As a novel result, we further found important population differences. We highlight that Blue Petrels did not have a northerly distribution or high trophic position, which usually account for elevated Hg concentrations in Southern Ocean seabirds. Instead, they have the most southern winter distribution of our three study species, and feed mainly on crustaceans, except on Kerguelen where myctophid fish constitute a substantial proportion of the diet. As other seabirds exposed to high Hg levels in winter, they have a notable temporal carry-over of high blood Hg values into the breeding season.

While the Kerguelen population of Blue Petrels has been tracked recently (e.g., Quillfeldt et al., 2015, 2020; Chérel et al., 2016), there are no diet or tracking data from the major population at Diego Ramírez. Our study suggests that this population has particularly high exposure to Hg (e.g., **Figure 3** and **Supplementary Figure 8**), which can be an additional stressor and impact reproduction and survival in birds (Goutte et al., 2014; Mills et al., 2020). Further study of their movements and foraging ecology are therefore required, in particular to confirm if the high Hg concentrations in feathers are related to differences in diet or sea-ice exposure. Additionally, a comparison of Hg in flight feathers, and of spatial and temporal variation in Hg concentrations of their crustacean and fish prey in relation to biogeography and ecology would help reveal the factors driving differences among seabird species in terms of Hg exposure and contamination. A combination of ship-based and tracking studies could address the question of how the foraging and movement ecology of predators and spatial differences interact to produce the patterns in Hg burdens observed in these and other wildlife in the Southern Ocean.

## DATA AVAILABILITY STATEMENT

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## ETHICS STATEMENT

The study involved wild individuals and was carried out under permits from the Falkland Islands Government (Environmental Planning: R21.2012), and the Animal Ethic Committee of the IPEV. All work conducted at Bird Island was approved by the

Ethics Committee of the British Antarctic Survey and carried out under permit from the Government of South Georgia and the South Sandwich Islands. Seabird work in Diego Ramírez was approved by Res. N° 959 and N° 6093, Servicio Agrícola y Ganadero (Agriculture and Livestock Service), Chile.

## AUTHOR CONTRIBUTIONS

PQ and PB conceived and designed the study. PQ, JM, JN, RP, and CS carried out the fieldwork. YC and PB carried out the lab work. PQ, JN, RP, KD, and YC carried out the data curation. PQ carried out the data analyses. PQ and PB drafted the manuscript. All authors reviewed the final draft of the manuscript.

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## SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fevo.2022.915199/full#supplementary-material>

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